



# **Water Resources Research**

# RESEARCH ARTICLE

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#### **Key Points:**

- Variations in mean annual ET<sub>N</sub>/ ET<sub>D</sub> ratio (5.1%–11.7%) across the sites were mostly determined by the spatial variability in ET<sub>D</sub>
- ET<sub>D</sub> and ET<sub>N</sub> responded differently to climatic drivers from halfhourly to annual scales, which also depended on ecosystems types
- At site levels, leading controls were Rn for half-hourly ET<sub>D</sub>, leaf area index (LAI), and soil water content (SWC) for daily to annual ET<sub>D</sub>, and SWC for half-hourly to annual ET<sub>N</sub>

#### **Supporting Information:**

Supporting Information may be found in the online version of this article.

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# Comparison of Nighttime With Daytime Evapotranspiration Responses to Environmental Controls Across Temporal Scales Along a Climate Gradient

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**Abstract** Understanding daytime  $(ET_D)$  and nighttime  $(ET_N)$  evapotranspiration is critical for accurately evaluating terrestrial water and carbon cycles. However, unlike  $ET_D$ , the factors influencing  $ET_N$  remain poorly understood. Here, long-term  $ET_D$  and  $ET_N$  data from five FLUXNET sites along a climate gradient in Northern Australia were analyzed to compare their responses to environmental drivers at different temporal scales. We found that (a) across the sites, mean annual  $ET_N/ET_D$  ( $\overline{ET_N}$  ) ranged between 5.1% and 11.7%, which was mainly determined by  $\overline{ET_D}$  variations. Particularly, vegetation and meteorological variables mostly controlled  $\overline{ET_D}$ , while  $\overline{ET_N}$  was largely related to air temperature and net radiation (Rn) due to lower nighttime atmospheric water demands; (b) At individual sites,  $ET_D$  and ET<sub>N</sub> exhibited higher correlations with meteorological and vegetation variables at monthly timescales than at annual timescales. Monthly  $ET_D$  and  $ET_N$  were also strongly coupled, especially under drier climatic conditions. At daily timescales, leaf area index and soil water content (SWC) controlled  $ET_D$ with SWC being more important at drier sites; whereas, SWC was the dominant factor controlling  $ET_N$ . At half-hourly timescales, the boosted regression tree method quantitively showed that  $ET_D$  and  $ET_N$ were controlled by Rn and SWC, respectively. Overall, the results showed that  $ET_N$  was less responsive to environmental variables, illustrating that  $ET_D$  and  $ET_N$  responded differently to diverse climate regimes and ecosystems at varying temporal scales. These findings provide a critical evaluation for contrasting  $ET_D$  and  $ET_N$  interactions in constantly changing environments, which has important implications for ecosystem water balance and land surface processes modeling.

## 1. Introduction

As one of the key components in terrestrial water cycles, actual evapotranspiration (ET) links water, energy, and carbon cycles via plant physiological activities (Fisher et al., 2017; Katul et al., 2012). Although ET is primarily composed of daytime ET ( $ET_D$ ), a growing body of observational and modeling evidence suggests that nighttime ET ( $ET_D$ ) is also important for various ecohydrological and physiological processes from local to global scales (Forster, 2014; Padrón et al., 2020; Whitley et al., 2013). Globally, Padrón et al. (2020) showed that the average  $ET_N/ET$  value was about 6.3% when calculated from the FLUXNET2015 data set and 7.9% from simulation results of global models. Moreover,  $ET_N$  can exert significant impacts on the estimation of global water and carbon budgets, particularly in arid and semiarid regions (Lombardozzi et al., 2017; Resco De Dios et al., 2015), since the consideration of nighttime water use as indicated by  $ET_N$  can noticeably reduce water use efficiency (WUE, defined as the ratio of  $CO_2$  assimilation rate over water vapor loss) (Chaves et al., 2016). At field scales, neglecting  $ET_N$  can cause the underestimation of total daily ET and the incomplete closure of energy balance, thus resulting in underestimated water requirements for crops (Skaggs & Irmak, 2011). Therefore, in addition to the role of  $ET_D$ , a better understanding of nighttime water loss from terrestrial ecosystems also bears many important implications for hydrological, climatic, and agricultural studies (Agam et al., 2012; Berkelhammer et al., 2013; Wang & Dickinson, 2012).

To accurately evaluate the impacts of  $ET_D$  and  $ET_N$  on ecohydrological and biogeochemical processes, it is critical to elucidate how they respond to changes in environmental conditions. However, the majority of relevant nighttime studies focused on nighttime transpiration ( $Tr_N$ ), which is part of  $ET_N$  and can account

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for up to 69% of total daily transpiration (Tr) for certain equatorial species (Forster, 2014). The mechanism that controls  $Tr_N$  is mainly attributed to incomplete stomatal closure at night (Dawson et al., 2007; Duursma et al., 2019), although other regulation mechanisms have also been proposed, such as the circadian regulation of nighttime water use by plants (Resco De Dios et al., 2015). Despite the debate on the underlying regulation mechanisms, a wealth of studies have revealed the complex interactions of  $Tr_N$  with surrounding environments (Iritz & Lindroth, 1994; Moore et al., 2008). The spatiotemporal variations in  $Tr_N$  have been linked to changes in soil water content (SWC) (Barbeta et al., 2012; O'keefe & Nippert, 2018), vapor pressure deficit (VPD) (Bucci et al., 2004; Tolk et al., 2006), and wind speed (WS) (Karpul & West, 2016). However, inconsistent findings have also been reported. For instance,  $Tr_N$  was shown to be positively related to VPD and SWC (Dawson et al., 2007; Moore et al., 2008), while no or negative relationships were also found between  $Tr_N$  and VPD (or SWC) (Barbour & Buckley, 2007; Phillips et al., 2010; Resco De Dios et al., 2015). In addition, biotic factors (e.g., plant species and ecosystem type) might also play important roles in determining  $Tr_N$  (O'keefe & Nippert, 2018; Zeppel et al., 2014). As a result of the combined influence of biotic and abiotic factors,  $Tr_N$  (and thus  $ET_N$ ) is highly variable among different plant species, ecosystems, and climate regimes (see Table S1 in the supplemental information for details).

Although previous work has focused on  $Tr_N$ , it is more relevant from hydrological and water resources management perspectives to quantify  $ET_N$  and particularly how it changes with environmental variables. As different components of ET (e.g., Tr, and soil and canopy evaporation) respond to environmental factors differently (Tsuruta et al., 2016), it would be questionable to directly extrapolate the findings obtained from  $Tr_N$  studies to understand  $ET_N$  behaviors under changing environmental conditions. Similar to  $Tr_N$ ,  $ET_N$  can be affected by a number of biotic and abiotic factors. At field scales, Skaggs and Irmak (2011) showed that  $ET_N$  was a function of weather conditions (e.g., VPD, WS, and net radiation-Rn), and the relative influence of these variables further varied with crop growing stages. Using eddy-covariance (EC) records from two forested ecosystems, Novick et al. (2009) observed that  $ET_N$  was primarily driven by WS and VPD, and  $Tr_N$  was the leading contribution to  $ET_N$ . By contrast, based on lysimeter measurements under natural grass covers, Groh et al. (2019) found that  $ET_N$  was mainly composed of evaporation and positively correlated with WS. The authors also showed the importance of night dew formation in controlling  $ET_N$  at the study sites. In short, the dominant environmental factors controlling  $ET_N$  were inconsistent in previous studies and remained poorly understood (Groh et al., 2019; Padrón et al., 2020; Zeppel et al., 2014).

More importantly, previous studies showed that dominant controls on  $ET_D$  and  $ET_N$  might be different (Barbeta et al., 2012; Karpul & West, 2016; O'keefe & Nippert, 2018), implying that  $ET_N$  cannot simply be estimated from  $ET_D$  alone and should be investigated separately (Groh et al., 2019; Ogle et al., 2012). For example, Groh et al. (2019) showed that at grassland sites from Germany, Rn and VPD were the dominant controls on  $ET_D$ , while  $ET_N$  was more strongly influenced by WS. O'keefe and Nippert (2018) argued that daytime and nighttime stomatal conductance should be parameterized separately in land surface models. In addition, the main controls on ET can change across temporal scales. For instance, Scott and Biederman (2019) found that the dominant controls on ET changed from subdaily (weather) to seasonal (SWC) scales in a semi-arid savanna ecosystem, indicating that controlling factors of  $ET_D$  and  $ET_N$  might also vary with temporal scales. Part of the reason can be attributed to the varying correlations among environmental factors across temporal scales, as stressed by Liu et al. (2020).

For the global study, Padrón et al. (2020) showed that for most FLUXNET sites, there were positive correlations of daily  $ET_N/ET$  with local air temperature  $(T_a)$ , Rn, and SWC at individual sites; however, no clear relationships of mean annual  $ET_N/ET$  ( $\overline{ET_N}/\overline{ET}$ , where overbar denotes mean annual timescales hereinafter) with precipitation  $(\overline{P})$  and  $\overline{T_a}$  were found across those sites. With regards to biotic factors, the relationship between  $\overline{ET_N}/\overline{ET}$  and ecosystem types was also weak across FLUXNET sites (Padrón et al., 2020). As ET is primarily composed of  $ET_D$ , the lack of general patterns from the study of Padrón et al. (2020) clearly demonstrated the poor correlation between  $ET_D$  and  $ET_N$  across different climate regimes and ecosystem types. Based on above discussions, we hypothesize that this could be attributed to the different responses of  $ET_D$  and  $ET_N$  to diverse climate regimes and ecosystems at varying temporal scales. Elucidation of the interactions of ET with environmental drivers at different timescales is critical for understanding a variety of ecohydrological and land surface processes. For instance, annual and monthly ET processes are related to vegetation phenology and therefore important for estimating terrestrial productivity (Berkelhammer

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et al., 2013). At daily and subdaily timescales, ET dynamics involving areas from genetics to physiology would affect plant performance and WUE (Chaves et al., 2016). Thus, the aforementioned studies underscore the necessity of differentiating the mechanistic responses of  $ET_D$  and  $ET_N$  to influencing factors at different temporal scales, for example, for quantifying and/or modeling  $ET_D$  and  $ET_N$  purposes.

To test our hypothesis, the spatiotemporal characteristics of  $ET_D$  and  $ET_N$  as well as their relationships with environmental drivers were analyzed using long-term EC records obtained at five FLUXNET sites along a climate gradient from Northern Australia. Based on data from the Australian Bureau of Meteorology (data available from 1973 to 2015), northern Australia exhibits a strong P gradient along the north-south direction with a monsoonal climate of distinct wet and dry seasons (Beringer et al., 2016). This P gradient aligns with structural and floristic changes in the composition of local vegetation communities (Hutley et al., 2011; Williams et al., 1996), making this region suitable for studying the spatiotemporal variations in  $ET_D$  and  $ET_N$  and for examining their responses to environmental variables. The responses of  $ET_D$  and  $ET_N$  to environmental controls along this climate gradient were analyzed from mean annual to half-hourly timescales. In addition, a nonlinear machine learning method, namely the boosted regression tree (BRT), was employed to quantify the relative controls of environmental factors on  $ET_D$  and  $ET_N$  at half-hourly timescales.

#### 2. Materials and Methods

#### 2.1. Site Description

For the purpose of this study, five sites installed with EC towers were selected (Table 1), which are distributed along the North Australian Tropical Transect (NATT; Figure 1). The NATT spans roughly 900 km, stretching from Darwin (about 12°S) in the mesic north of the Northern Territory to the xeric edge of the Tanami Desert (about 17°S) (Schulze et al., 1998). Along the NATT,  $\overline{T_a}$  shows no significant gradient (~26°C across the sites), while  $\overline{P}$  decreases noticeably from ~1,800 mm/year in the north to ~200 mm/year in the south (Schulze et al., 1998). There is a pronounced climate seasonality in the study region, with a dry season extending from May to September.

Dominant vegetation communities in the study region tend to be savannas, which are characterized by a patchy woody overstory (mainly evergreen Eucalyptus and Corymbia tree species) and an understory (mainly C4 grasses) (Whitley et al., 2016). The abundance and height of the overstory eucalypts decrease with declining  $\bar{P}$  from north to south, and understory grasses die off immediately at the end of wet seasons (Hutley et al., 2011). In addition, tropical grasslands dominate the intermediate, semi-arid section of the NATT (e.g., at the AU-DaP and AU-Stp sites [Site ID from the FLUXNET]). Soil types range from sand-dominated Kandosols in the north to Vertosols or cracking clays that are more widespread in the south (Australian Soil Classification [Isbell, 2002]). The alternating wet and dry seasons and the redox of metal-rich soils produce duricrusts in soils with varying depths, which ultimately limit vegetation growth (Dilshad et al., 1996; Hutley et al., 2011). Relevant information for each site is summarized in Table 1.

# 2.2. Data Acquisition

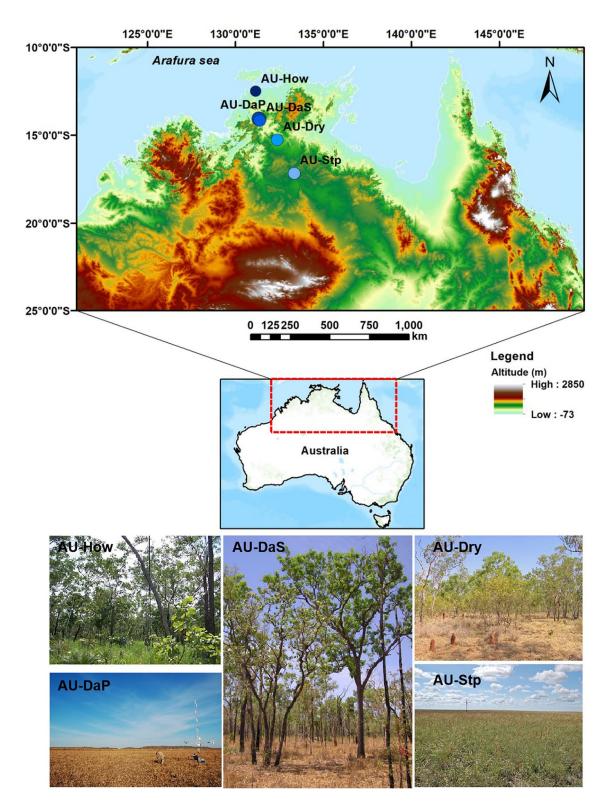
The FLUXNET is a global network equipped with EC towers to measure site-level water, carbon, and energy exchanges between the atmosphere and biosphere (Baldocchi et al., 2001). In this study, half-hourly meteorological and soil data (e.g., latent heat-*LE*, sensible heat-*H*, soil heat flux-*G*, *P*, SWC, *Ta*, *Rn*, VPD, WS, and friction velocity-*u\**) measured at the five FLUXNET sites were retrieved from the FLUXNET 2015 Tier 1 data set (https://fluxnet.org/data/fluxnet2015-dataset/). Here, the data set is briefly discussed, and the detailed descriptions can be found at http://www.ozflux.org.au/and in Beringer et al. (2016). Specifically, a CSAT3 sonic anemometer (Campbell Scientific, Logan, Utah, USA) and a Li-7500 infra-red gas analyzer (LI-COR, Lincoln, Nebraska, USA) are mounted at the top of the EC towers with the height of 23 m, 15 m, 23 m, 15 m, and 4.8 m above the ground at AU-How, AU-DaP, AU-DaS, AU-Dry, and AU-Stp, respectively. The EC towers record wind field (WS and *u\**) and H<sub>2</sub>O concentrations at 10 Hz or 20 Hz, which are then integrated to half-hourly data. Radiation (NR-Lite, Kipp & Zonen, Delft, Netherlands), and *Ta* and relative humidity (HMP45AC or HMP45 A, Vaisala, Helsinki, Finland) are measured at the tower top, while *P* (TB3, HyQuest Solutions, New South Wales, Australia; CS702, Campbell Scientific, Utah, USA) is measured at the ground. Soil variables, including SWC (CS616, Campbell Scientific, Utah, USA) and *G* (HFT3, Campbell Scientific,

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	Canopy height (m)	15	0.3	16.4	12.3	0.5
	Species	Eucalyptus miniata Eucalyptus tetrodonta	Chamaecrista rotundifolia Digitaria milijiana	Eucalyptus tetrodonta Corymbia latifolia	Eucalyptus tetrodonta Corymbia terminalis	Grassland Mitchell Grass
	Ecosystem	Savanna	Grassland	Savanna	Savanna	Grassland
	Soil type	Sandy loam (kandosol)	Sandy loam (kandosol)	Sandy loam (kandosol)	Sandy loam (kandosol)	Cracking clay (vertisol)
	$\frac{\mathrm{Daytime}}{Tr/\overline{ET}}$	0.56	0.64	0.51	0.50	0.46
	$\overline{ET_p}/\overline{P}$	0.98	1.21	1.32	2.07	2.82
	$\overline{ET_p}$ (mm/year) $\overline{ET_p}/\overline{P}$	1689.5	1704.2	1851.6	1843.6	2146.2
nsect	$\overline{T_a}(^{\circ}C/\text{year})$ $\overline{P}(\text{mm/year})$	1715.5 (1112.0– 2287.4) <sup>a</sup>	1402.8 (945.1–1723.9)	1400.1 (1167.7–1599.4)	890.4 (655.1–1396.2)	761.1 (473.6–1135.6)
n Tropical Tra	$\overline{T_a}(^{\circ}C/\mathrm{year})$	26.7	25.5	26.7	26.9	26.1
th Australia	Elevation (m)	64	70	110	175	250
1long the Nor	Latitude Longitude Elevation	131.15°E	131.32°E	131.39°E	132.37°E	133.35°E
tudy Sites	Latitude (°)	12.49°S	14.06°S	14.16°S	15.26°S	17.15°S
Table 1 Information for the Five Study Sites Along the North Australian Tropical Transect	year	2002–2014 12.49°S	2008–2013 14.06°S	2009-2014 14.16°S 131.39°E	2009–2014 15.26°S	2009–2014 17.15°S 133.35°E
Table 1Information	Site ID	AU-How	AU-DaP	AU-DaS	AU-Dry	AU-Stp

 ${}^{a}$ The range of annual precipitation during the study period;  $\overline{T_a}$ : mean annual temperature;  $\overline{P}$ : mean annual precipitation;  $\overline{ET_p}$ : mean annual precipitation;  $\overline{ET_p}$ : mean annual aridity index; Daytime  $\overline{Tr}/\overline{ET}$ : long-term ratio of transpiration to actual evapotranspiration during the daytime (following the method of Zhou et al., 2016 as described in the supplemental materials).

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 $\textbf{Figure 1.} \ \ Locations \ of the five FLUXNET sites used in this study along the North Australian Tropical Transect. Photos are shown to illustrate the differences in the structure of vegetation (http://www.ozflux.org.au/).$ 

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Utah, USA), are measured at the depth of 5 cm and/or 15 cm adjacent to the towers. To be consistent, SWC and *G* data at the depth of 5 and 15 cm, respectively, were used in this study.

To quantify vegetation conditions at the sites, leaf area index (LAI) data were retrieved from the Moderate Resolution Imaging Spectroradiometer (MODIS) MCD15A3H data set (starting in 2002; version 6; https://e4ftl01.cr.usgs.gov/MOTA/), based on the geographical coordinates of each site. The MCD15A3H data set has a temporal resolution of 4 days and a spatial resolution of 500 m, and therefore post-processing was needed to obtain a daily LAI data set. First, based on the quality control information, data with unfavorable conditions (e.g., the presence of cloud and/or snow cover) were filtered out. A cubic smoothing spline method was then used for interpolating the 4-day interval data into daily values (Horn & Schulz, 2010), and the Savitzky-Golay filter was used to smooth out noises in the LAI time series (Chen et al., 2004). At half-hourly timescales, the LAI data remained constant for the analysis during each calendar day.

#### 2.3. Processing of EC Data

The FLUXNET data set has been through rigorous data processing procedures with four main blocks: data quality assurance and control, the Energy & Water Fluxes processing block, the Carbon Fluxes processing block, and the Meteorological Variables processing block (Pastorello et al., 2020; Vuichard & Papale, 2015). The detailed information can be found at <a href="https://fluxnet.org/data/fluxnet2015-dataset/data-processing/">https://fluxnet.org/data/fluxnet2015-dataset/data-processing/</a>. Note that the original daytime and nighttime data values in the data set were separated based on the potential incoming shortwave radiation (the theoretical maximum radiation at the top of atmosphere), which was computed using latitude/longitude coordinates and time (Pastorello et al., 2020).

In this study, to ensure the suitability of the EC data for assessing  $ET_D$  and  $ET_N$ , additional data processing procedures were performed. First, to resolve the energy balance closure problem of EC measurements, the energy closure ratio (ECR) (defined as (LE + H)/(Rn-G)) was computed at annual timescales for each site (Liu et al., 2019). The resulting average annually ECR over the study periods was 0.93, 0.89, 0.91, 0.84, and 0.98 for the AU-How, AU-DaP, AU-DaS, AU-Dry, and AU-Stp site, respectively, which were in accordance with those (0.87  $\pm$  0.15) reported for the FLUXNET sites (Stoy et al., 2013); therefore, no correction was made for closing the energy balance in this study. Second, data records (AU-How: 2001–2014; AU-DaP: 2007–2013; AU-DaS: 2008–2014; AU-Dry: 2008–2014; AU-Stp: 2008–2014) in the first calendar year for each site were omitted (e.g., the year 2001 was removed for the AU-How site) due to the large amount of missing data. Third, LE data were removed if they were collected during stable atmospheric conditions at night according to  $u^*$  values. Site-dependent  $u^*$  thresholds were determined annually following Reichstein et al. (2005). After the  $u^*$  quality filtering, 25.1%, 35.6%, 26.9%, 27.1%, and 39.3% of the original nighttime data were retained for the AU-How, AU-DaP, AU-DaS, AU-Dry, and AU-Stp site, respectively. The percentage of the remaining data for each calendar year at the five FLUXNET sites is provided in Table S2.

To fill the missing nighttime LE data, the online tool (available at <a href="https://www.bgc-jena.mpg.de/bgi/index.php/Services/REddyProcWeb">https://www.bgc-jena.mpg.de/bgi/index.php/Services/REddyProcWeb</a>) was used. This online gap-filling tool was derived from the technique proposed by Falge et al. (2001), with the consideration of the temporal auto-correction of the fluxes and the co-variations of the fluxes with other meteorological variables (Reichstein et al., 2005). The performance of the gap-filling method was evaluated by comparing observed LE with filled LE according to Moffat et al. (2007), and the results are provided in Table S3. The half-hourly daytime LE (note no gap filling was performed for daytime LE) and gap-filled nighttime LE were used to calculate half-hourly  $ET_D$  and  $ET_N$  (here, negative values were replaced by zero). Half-hourly data (e.g.,  $ET_D$ ,  $ET_N$  and environmental variables at corresponding daytime or nighttime) were then summed up to obtain daily, monthly, and annual values. As dew formation might be important in some areas (Groh et al., 2019; Jacobs et al., 2006), the half-hourly gap-filled LE with negative values were also summed up to compute the amount of dew formation at the sites.

#### 2.4. Statistical Analysis

The BRT method developed by Elith et al. (2008) was applied to quantify how environmental factors influenced half-hourly  $ET_D$  and non-gap-filled  $ET_N$  (calculated from nongap-filled half-hourly nighttime LE) along the climate gradient of the NATT. As a nonlinear machine learning method, the BRT method differs

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from traditional regression methods in that it uses the boosting technique to adaptively combine large numbers of relatively simple tree models for predicting the relationship between predictor and response variables (Elith et al., 2008). Specifically, based on the recursive binary splitting method, tree-based models are fitted iteratively to the observations, until a minimum deviance is reached between observed and predicted data. The relative impact of a predictor variable on a response variable is quantified based on the number of times the predictor variable is used for splitting the trees, which is then weighted by the squared improvement to the whole tree-based model (Breiman et al., 1984). The BRT method does not require any specific data distributions, and thus is suitable for analyzing datasets with outliers, interactions among predictor variables, and missing data (Elith et al., 2008).

The open-source BRT package developed by Elith et al. (2008) was employed in this study using the R software (version 3.5.1, R Core Team, 2018), which requires four key input parameters (i.e., learning rate-lr, tree complexity-tc, number of trees-nt, and bag fraction-bg). The parameters lr and tc determine the contribution of individual trees to the growing model and the structure of a tree, respectively, which jointly determine nt for optimal predictions. The maximum and minimum nt values were set to be 30,000 and 1,000, respectively, as suggested by Elith et al. (2008). Moreover, the stochasticity of the boosted model is controlled through bg that specifies the proportion of data to be selected at each iteration from the full training set. To obtain optimal predictive results, combinations of parameter values (e.g., lr [0.003, 0.005, 0.01, 0.03, 0.05], tc [from 7 to 10 by an increment of 1], and bg [0.5, 0.75]) were tested (Liu et al., 2019), and an optimal parameter set, which provided the minimum predictive deviance between observed and predicted data, was chosen for each site to compute the relative contribution of each predictor variable in controlling  $ET_D$  and  $ET_N$ .

Following Elith et al. (2008), a 10-fold cross-validation method was used to build the BRT model and assess its performance. Data were first randomly partitioned into 10 subsets, with 9 subsets for training the model and the remaining one for testing the model performance. This procedure was repeated 10 times to get the predictive results. Note that the multicollinearity in predictor variables (e.g., Rn, SWC, LAI, VPD,  $T_a$ , and WS) may affect the performance of the BRT model for predicting response variables (e.g.,  $ET_D$  and  $ET_N$ ). To resolve this issue, Pearson's correlation coefficient (r) and variance inflation factor (VIF) analysis (i.e., |r| < 0.7 and VIF < 5) were used to test the multicollinearity among predictor variables (Dormann et al., 2013). The performance of the BRT model was evaluated using the Nash-Sutcliffe efficiency, root mean square error, and coefficient of determination by comparing observed and modeled half-hourly  $ET_D$  and  $ET_N$  for each site (Table S4).

# 3. Results and Discussion

# 3.1. Mean Annual $ET_D$ and $ET_N$ Patterns Along the Climate Gradient

#### 3.1.1. Spatial Variability in Mean Annual $ET_D$ and $ET_N$

Along the NATT from AU-How to AU-Stp,  $\overline{P}$  decreased from 1715.5 mm/year to 761.1 mm/year during the study periods, while mean annual potential evapotranspiration ( $\overline{ET_p}$ ) (calculated by the Penman-Monteith equation [Allen et al., 1998]) showed an opposite trend from 1689.5 mm/year to 2146.2 mm/year; accordingly, mean annual aridity index ( $\overline{ET_p}/\overline{P}$ ) varied from 0.98 to 2.82. The mean annual dew formation was about 4.5 mm/year at the sites, which was negligible compared to the values reported for humid grassland sites in Germany (Groh et al., 2019) and forested sites in central Colorado (Berkelhammer et al., 2013). Thus, the impact of dew formation was not considered in this study; however, it should be noted that the amount of dew formation was inferred based on the EC measurements, which might lead to uncertainties in the dew formation estimates (Moro et al., 2007). Nevertheless, the pronounced climate gradient along the NATT is ideal for diagnosing the responses of  $ET_D$  and  $ET_N$  to varying climatic conditions.

Among the five sites,  $\overline{ET}$  varied from 1158.2 mm/year at AU-How in the north to 609.7 mm/year at AU-Stp in the south, which expectedly was primarily composed of  $ET_D$  (Table 2). Intriguingly, it appeared that  $\overline{ET_D}$  exhibited a stronger variation than  $\overline{ET_N}$  along the NATT. For demonstration, Figures 2a–2l show the relationships of  $\overline{ET_D}$  and  $\overline{ET_N}$  with different environmental variables across the sites. In general, the correlations of  $\overline{ET_D}$  with meteorological and vegetation variables were notable, while  $\overline{ET_N}$  only showed good relationships with  $\overline{T_a}$  and  $\overline{Rn}$ , illustrating the contrasting responses of  $\overline{ET_D}$  and  $\overline{ET_N}$  to the environmental

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**Table 2**Mean Annual Actual Evapotranspiration ( $\overline{ET}$ ), Daytime Evapotranspiration ( $\overline{ET_D}$ ), Nighttime Evapotranspiration ( $\overline{ET_D}$ ), and  $ET_N/ET_D$  ( $\overline{ET_N}$ ) for the Five Study Sites Along the North Australian Tropical Transect

Site ID	ET(mm/year)	$\overline{ET_D}$ (mm/year)	$\overline{ET_N}$ (mm/year)	$\overline{ET_N}$ / $\overline{ET_D}$ (%)
AU-How	1158.2 (1004.5–1413.4) <sup>a</sup>	1090.9 (928.2–1319.8)	67.2 (45.8–93.6)	6.2 (4.6-8.2)
AU-DaP	749.3 (611.4–915.8)	712.1 (592.9–861.6)	37.3 (18.6-54.2)	5.1 (2.8-6.6)
AU-DaS	930.9 (882.8–972.6)	849.0 (819.1-882.7)	81.9 (63.6-91.3)	9.6 (7.8–10.8)
AU-Dry	884.4 (804.3-1164.4)	802.5 (728.7–1069.3)	81.9 (74.0-95.2)	10.3 (8.9–12.3)
AU-Stp	609.7 (467.3-854.3)	546.4 (413.7-771.9)	63.2 (40.3-82.5)	11.7 (7.9–14.2)

<sup>&</sup>lt;sup>a</sup>The range of annual variables during the study period.

gradient. In particular, the correlation between  $\overline{ET_N}$  and  $\overline{P}$  was significantly lower than that between  $\overline{ET_D}$  and  $\overline{P}$  (Figures 2b and 2h). This suggests that as atmospheric demands for ET became lower at night, the constraint of water supplies (indicated by  $\overline{P}$  levels) on  $ET_N$  weakened. Instead,  $\overline{ET_N}$  across the sites was dominantly controlled by  $\overline{T_a}$  (most likely through affecting soil evaporation; Liu et al., 2015). As such, the narrow  $\overline{T_a}$  range led to a relatively weaker variation in  $\overline{ET_N}$  along the NATT.

With higher atmospheric demands for  $ET_D$  (especially due to enhanced daytime plant Tr; see Table 1 for daytime  $\overline{Tr}/\overline{ET}$ ), the impacts of vegetation (e.g.,  $\overline{LAI}$ ) and relevant meteorological variables (e.g.,  $\overline{VPD}$ ) on  $\overline{ET_D}$  emerged across the sites. These factors with large variabilities led to a stronger variation in  $\overline{ET_D}$  along the NATT. Nevertheless, the data along the NATT provided strong observational evidence that  $\overline{ET_D}$  and  $\overline{ET_N}$  could respond differently to the climate gradient with  $\overline{ET_D}$  showing more distinct interactions with influencing factors, primarily due to the increased daytime plant Tr.

# 3.1.2. Environmental Controls on Mean Annual $ET_N/ET_D$ Ratios

The ratio of  $ET_N/ET_D$  (or  $ET_N/ET$ ) is an important metric for quantifying the relative importance of nighttime water loss over terrestrial ecosystems (Dawson et al., 2007; Groh et al., 2019; Iritz & Lindroth, 1994; Novick et al., 2009; Padrón et al., 2020). With the pronounced climate and vegetation gradients along the NATT, the  $\overline{ET_N}/\overline{ET_D}$  ratio varied noticeably among the sites (Table 2), which were comparable to previous studies. For instance,  $Tr_N$  was 5%–15% of daytime Tr in a eucalypt woodland in Western Australia (Dawson et al., 2007); grass and tree ecosystems in the Southeastern USA exhibited annual  $ET_N/ET_D$  of 8%–9% using EC measurements (Novick et al., 2009). The relationships of  $\overline{ET_N}/\overline{ET_D}$  with different environmental variables are presented in Figures 2m-2r. Interestingly, although  $T_a$  and Rn were the leading controls on  $\overline{ET_N}$  as shown above, their relationships with  $\overline{ET_N}/\overline{ET_D}$  deteriorated significantly. Instead, the strong correlations of  $\overline{ET_N}/\overline{ET_D}$  with  $\overline{P}$ ,  $\overline{VPD}$ , and  $\overline{LAI}$  mainly reflected the spatial variation in  $\overline{ET_D}$  along the NATT. By contrast, Zeppel et al. (2010) reported that  $Tr_N$  mostly determined the seasonal variation in  $ET_N/ET_D$  in a eucalypt-dominated woodland in New South Wales, Australia. Thus, it is reasonable to postulate that the contrasts in  $\overline{ET_D}$  and  $\overline{ET_N}$  responses to climatic drivers, which can be further complicated by vegetation conditions, can greatly mask the relationships of  $\overline{ET_N}/\overline{ET_D}$  (or  $\overline{ET_N}/\overline{ET}$ ) with  $\overline{P}$  and  $\overline{T_a}$  among different climate regimes and ecosystems, making it difficult to seek a general pattern of  $\overline{ET_N}/\overline{ET}$  with environmental variables at the global scale as demonstrated by Padrón et al. (2020).

# 3.2. Monthly and Annual Variations in $ET_D$ and $ET_N$

Monthly  $ET_D$  and  $ET_N$  showed clear seasonal patterns with higher values occurring in wet seasons (Figures 3a and 3b). Note that Groh et al. (2019) reported higher  $ET_N$  during nongrowing seasons, mostly due to longer nocturnal hours in winter times at the authors' study sites. In contrast, the nocturnal hours at the sites in Northern Australia are similar among different seasons. Moreover, with the pronounced climate seasonality, the temporal variations in monthly  $ET_D$  and  $ET_N$  at site levels showed good correlations with

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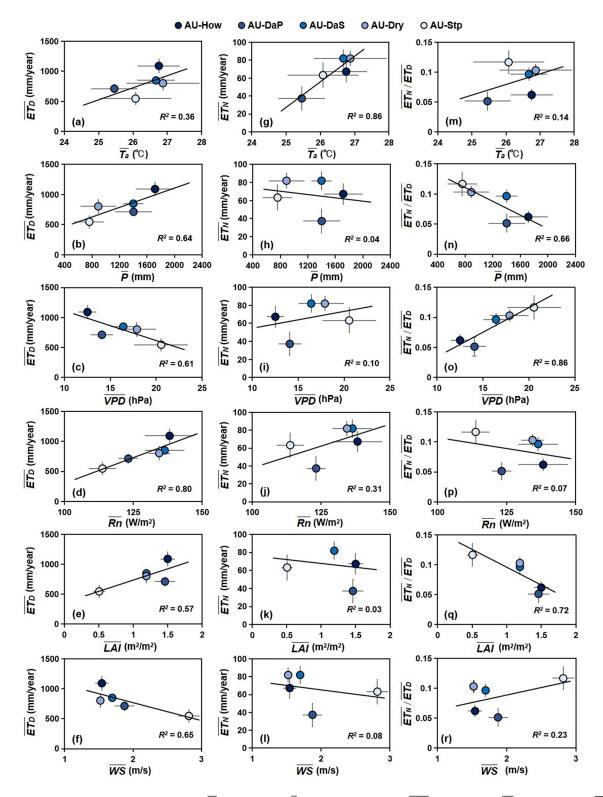


Figure 2. Relationships of mean annual air temperature  $(\overline{T_a})$ , precipitation  $(\overline{P})$ , vapor pressure deficit  $(\overline{VPD})$ , net radiation  $(\overline{Rn})$ , leaf area index  $(\overline{LAI})$ , and wind speed  $(\overline{WS})$  with mean annual daytime and nighttime evapotranspiration  $(\overline{ET_D}, \overline{ET_N})$  for the five study sites encoded by gradient colors (annual mean  $\pm$  standard deviation).

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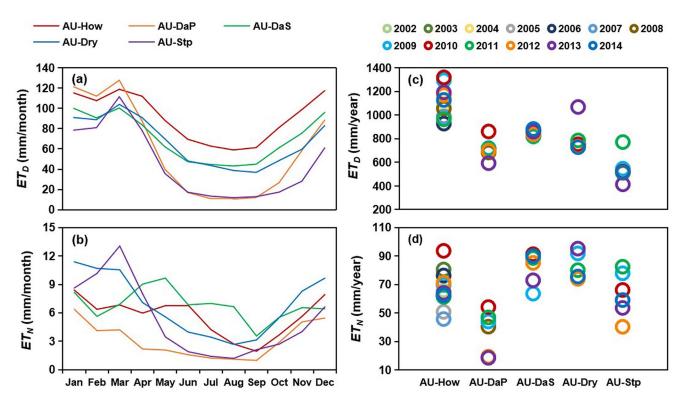


Figure 3. Average monthly daytime and nighttime evapotranspiration  $(ET_D, ET_N)$  and interannual variations in  $ET_D$  and  $ET_N$  for the five study sites.

meteorological and vegetation variables (Table 3; except for  $ET_N$  at AU-DaS). In particular, LAI and SWC were the leading factors of  $ET_D$ , and  $ET_N$  dynamics were dominated by SWC. Note that although  $\overline{T_a}$  largely determined the spatial variability in  $\overline{ET_N}$  along the NATT, local  $T_a$  became relatively less important in controlling monthly  $ET_N$  variations at each individual site.

Monthly  $ET_D$  and  $ET_N$  at each site were positively correlated with each other (Figure 4), which was in line with previous findings (Cavenderbares et al., 2007; Novick et al., 2009). Interestingly, the correlation between monthly  $ET_D$  and  $ET_N$  became progressively stronger with increasing climate dryness across the sites (as indicated by aridity index; except for AU-DaS). It suggested a stronger coupling between monthly  $ET_D$  and  $ET_N$  under drier climatic conditions, as monthly  $ET_D$  and  $ET_N$  showed more similar responses to environmental drivers at drier sites (Table 3). Recent investigation revealed that higher daytime photosynthetic rates generally led to higher  $Tr_N$ , due to the endogenous circadian rhythm of plant physiological activities (O'keefe & Nippert, 2018; Resco De Dios et al., 2015). To some degree, the strong correlations between monthly  $ET_D$  and  $ET_N$  in our results (Figures 3a, 3b and 4) supported this claim, which might be also affected by climatic conditions.

Interannual variability in ET is a longstanding research question as it is related to the critical roles of climate interannual variability in controlling annual variations in hydrological processes and biomass production (Fatichi & Ivanov, 2014; Milly & Dunne, 2002). As shown in Table 3, no strong correlations could be found to indicate the dominant factors driving the interannual variations in  $ET_D$  and  $ET_N$  at the study sites, although it is interesting to note that annual  $ET_N$  appeared to show relatively larger year-to-year variations than  $ET_D$  (Figures 3c and 3d). In general, LAI and SWC played important roles in determining  $ET_D$  variations, while the influencing factors of  $ET_N$  depended on ecosystem types, SWC for grass sites and meteorological variables (i.e., Rn, VPD, and  $T_A$ ) for savanna sites. Compared to the monthly data, the results suggested the decoupling of  $ET_D$  and  $ET_N$  with environmental variables at longer timescales (Table 3); however, it might also be attributed to the fact that the length of data records might not be long enough to reveal annual patterns of  $ET_D$  and  $ET_N$  with those variables. It is likely that different responses of  $ET_D$  and  $ET_N$  to environmental

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**Table 3**Spearman Correlation of Daytime and Nighttime Evapotranspiration  $(ET_D, ET_N)$  With Air Temperature  $(T_a)$ , Vapor Pressure Deficit (VPD), Soil Water Content (SWC), Net Radiation (Rn), Leaf Area Index (LAI), and Wind Speed (WS) for the Five Study Sites at Different Timescales Along the North Australian Tropical Transect

		Daily		Monthly		Annual	
Site ID	Variables	$ET_D$	$ET_N$	$ET_D$	$ET_N$	$ET_D$	$ET_N$
AU-How	$T_a$	0.37 <sup>a</sup>	0.13 <sup>a</sup>	0.29 <sup>a</sup>	0.18 <sup>b</sup>	0.15	0.25
	VPD	$-0.29^{a}$	$-0.14^{a}$	$-0.65^{a}$	$-0.36^{a}$	-0.18	0.22
	SWC	0.53 <sup>a</sup>	0.43 <sup>a</sup>	0.71 <sup>a</sup>	0.51 <sup>a</sup>	-0.05	-0.25
	Rn	0.45 <sup>a</sup>	0.23 <sup>a</sup>	0.20 <sup>b</sup>	0.35 <sup>a</sup>	0.27	0.67 <sup>t</sup>
	LAI	0.55 <sup>a</sup>	0.23 <sup>a</sup>	0.62 <sup>a</sup>	0.34 <sup>a</sup>	0.48	0.16
	WS	$-0.14^{a}$	0.19 <mark>a</mark>	$-0.38^{a}$	0.17 <sup>b</sup>	-0.005	0.23
AU-DaP	$T_a$	0.14 <sup>a</sup>	0.33 <sup>a</sup>	0.13	0.38 <sup>a</sup>	-0.14	-0.20
	VPD	$-0.56^{a}$	$-0.35^{a}$	$-0.72^{a}$	$-0.51^{a}$	-0.26	-0.54
	SWC	0.75 <sup>a</sup>	0.62 <mark>a</mark>	0.81 <sup>a</sup>	0.69 <sup>a</sup>	0.26	0.66
	Rn	0.42 <sup>a</sup>	0.38 <sup>a</sup>	0.26 <sup>b</sup>	0.48 <sup>a</sup>	0.20	-0.37
	LAI	0.80 <sup>a</sup>	0.41 <sup>a</sup>	0.84 <sup>a</sup>	0.47 <sup>a</sup>	0.31	0.31
	WS	$-0.27^{a}$	0.06 <sup>a</sup>	$-0.44^{a}$	0.003	0.6	-0.09
AU-DaS	$T_a$	0.40 <sup>a</sup>	$-0.20^{a}$	0.39 <sup>a</sup>	$-0.24^{b}$	0.14	0.49
	VPD	$-0.26^{a}$	-0.04	$-0.57^{a}$	-0.21	-0.03	-0.14
	SWC	0.63 <sup>a</sup>	0.21 <sup>a</sup>	0.77 <sup>a</sup>	0.25 <sup>b</sup>	-0.09	0.37
	Rn	0.63 <sup>a</sup>	$-0.07^{a}$	0.66 <sup>a</sup>	0.03	0.54	0.26
	LAI	0.64 <sup>a</sup>	$-0.13^{a}$	0.80 <sup>a</sup>	-0.12	0.81 <sup>b</sup>	0.35
	WS	$-0.30^{a}$	0.11 <sup>a</sup>	$-0.61^{a}$	-0.03	-0.20	-0.60
AU-Dry	$T_a$	0.09 <sup>a</sup>	0.27 <sup>a</sup>	0.07	0.34 <sup>a</sup>	0.09	0.20
	VPD	$-0.34^{a}$	$-0.41^{a}$	$-0.53^{a}$	$-0.55^{a}$	-0.43	0.54
	SWC	0.74 <sup>a</sup>	0.73 <sup>a</sup>	0.79 <sup>a</sup>	0.81 <sup>a</sup>	0.60	0.43
	Rn	0.40 <sup>a</sup>	0.51 <sup>a</sup>	0.46 <sup>a</sup>	0.74 <sup>a</sup>	0.09	0.26
	LAI	0.58 <sup>a</sup>	0.34 <sup>a</sup>	0.68 <sup>a</sup>	0.47 <sup>a</sup>	-0.09	0.06
	WS	$-0.37^{a}$	$-0.14^{a}$	$-0.72^{a}$	$-0.39^{b}$	-0.71	0.60
AU-Stp	$T_a$	0.16 <sup>a</sup>	0.27 <sup>a</sup>	0.22	0.45 <sup>a</sup>	-0.49	-0.26
	VPD	$-0.29^{a}$	$-0.44^{a}$	$-0.40^{a}$	$-0.43^{a}$	-0.66	-0.26
	SWC	0.71 <sup>a</sup>	0.63 <sup>a</sup>	0.79 <sup>a</sup>	0.76 <sup>a</sup>	0.89 <sup>b</sup>	0.77
	Rn	0.51 <sup>a</sup>	0.50 <sup>a</sup>	0.71 <sup>a</sup>	0.79 <sup>a</sup>	-0.09	0.26
	LAI	0.65 <sup>a</sup>	0.42 <sup>a</sup>	0.70 <sup>a</sup>	0.49 <sup>a</sup>	0.70	0.55
	WS	$-0.11^{a}$	-0.01	$-0.38^{a}$	$-0.26^{b}$	0.49	0.37

<sup>a</sup>Correlation is significant at the 0.01 level (two-tailed). <sup>b</sup>Correlation is significant at the 0.05 level (two-tailed).

drivers at site levels led to the varying degrees of the coupling strength between  $ET_D$  and  $ET_N$  at different temporal scales (monthly vs. annual timescales).

# 3.3. $ET_D$ and $ET_N$ Behaviors With Environmental Drivers at Daily Timescales

The relationships of daily  $ET_D$  and  $ET_N$  with environmental variables show similar patterns for the five sites (Table 3). During daytime, both LAI and SWC were the leading controls on daily  $ET_D$  with SWC being more important at drier sites (i.e., the AU-Dry and AU-Stp sites). Similar conclusions have been reached in previous studies (Han et al., 2020; Williams & Torn, 2015). By comparison, the correlations of LAI with daily  $ET_N$  considerably weakened, while the correlations between SWC and daily  $ET_N$  also weakened, but to lesser

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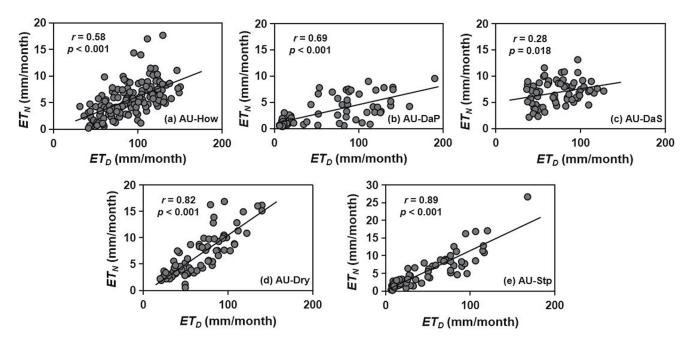


Figure 4. Relationships between monthly daytime evapotranspiration  $(ET_D)$  and nighttime evapotranspiration  $(ET_N)$  for the five study sites. r is the Pearson's correlation coefficient.

degrees. However, as SWC and VPD at daily timescales were strongly coupled (Figure S1), it is possible that the strong correlations of SWC with daily  $ET_D$  and  $ET_N$  were the byproduct of the impact of VPD on daily ET. Therefore, the responses of daily  $ET_D$  and  $ET_N$  to SWC and VPD require further investigation.

The question as to the relative importance of SWC and VPD in determining terrestrial land surface processes, such as ET and biomass production, is an ongoing debate in climate change studies (Liu et al., 2020; Novick et al., 2016; Sulman et al., 2016). To address this issue, Figure 5 shows the relationship of  $ET_D$  or  $ET_N$  with daytime or nighttime SWC ( $SWC_D$  and  $SWC_N$ , respectively) at the study sites. Overall, below certain SWC thresholds,  $ET_D^*$  was enhanced with increasing  $SWC_D$ , due to the constraint of water supplies on ET; whereas,  $SWC_D$  started to show a damping effect on  $ET_D^*$  when the thresholds were crossed, implying that factors other than  $SWC_D$  primarily affected  $ET_D$  (e.g., VPD; see the following analysis). With less atmospheric demands for  $ET_N$ , the threshold behavior of  $ET_N^*$  with  $SWC_N$  became less pronounced, suggesting that daily  $ET_N$  was less responsive to the change in SWC levels.

The threshold, at which daily ET starts to decrease with increasing SWC, is determined by a suite of variables, such as soil properties, meteorological conditions, and vegetation characteristics (Haghighi et al., 2018; Vivoni et al., 2008). For instance, there was no clear threshold at AU-Stp (Figure 5e). With the driest climate and highest atmospheric water demands, SWC was the primary factor controlling ET. By comparison, the threshold was smallest at AU-DaS among the five sites (Figure 5c), likely due to the coarse soil texture with a low water holding capacity as indicated by the lowest average level of SWC at this site (e.g., the sand fraction was 63% by weight according to the soil database of Shangguan et al., 2014). It is worth noting that with  $\sim$ 13.4 km apart, the AU-DaS and AU-DaP sites received similar  $\overline{P}$ ; however, both  $\overline{ET}_D$  and  $\overline{ET}_N$  were substantially higher at AU-DaS with lower SWC levels (Table 2), where deep-rooted trees could access deeper soil moisture for ET (as indicated by the canopy height in Table 1).

The above results underscored the necessity of combining the effects of SWC and other influencing factors for elucidating daily  $ET_D$  and  $ET_N$  dynamics (e.g., explaining the damping effect of SWC on ET). For simplicity, all observed data from the five sites were synthesized together to show the combined impact of SWC and VPD on daily  $ET_D$  and  $ET_N$  in Figure 6. At low daytime VPD ( $VPD_D$ ; e.g., below ~15 hPa), daily  $ET_D$  tended to increase rapidly with increasing  $VPD_D$ , as VPD can act as a driving force for stomatal water fluxes (Gu et al., 2006); whereas, their relationship became negative when  $VPD_D$  exceeded a critical value, due to stomatal regulation and the limitation of SWC (Gu et al., 2006; Leuning, 1995; Song et al., 2020). Note

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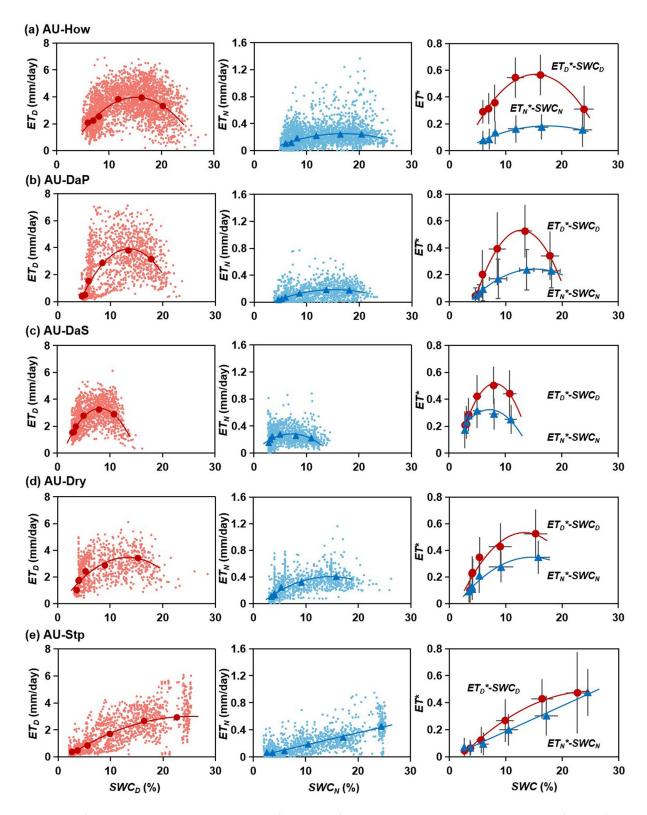
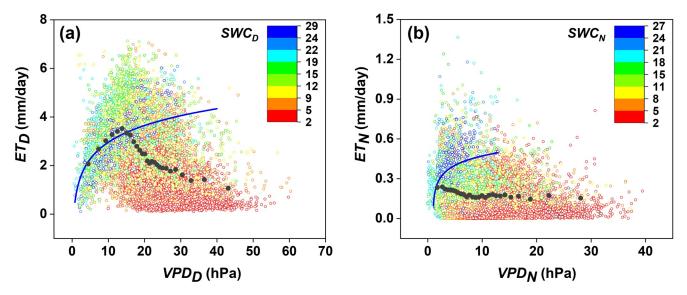


Figure 5. Relationships of daytime and nighttime soil water content  $(SWC_D, SWC_N)$  with daytime and nighttime evapotranspiration  $(ET_D, ET_N)$  at daily timescales. The data were binned according to the SWC range. Normalized daily  $ET_D$  and  $ET_N (ET_D^*, ET_N^*)$ ; that is,  $(ET - ET_{\min})/(ET_{\max} - ET_{\min})$ , where  $ET_{\max}$  and  $ET_{\min}$  are the maximum and minimum observed  $ET_D$  or  $ET_N$  during the study periods, respectively) are also given here. Polynomial curves are fitted through the average points, and horizontal and vertical bars represent standard deviations of SWC and  $ET^*$ , respectively.

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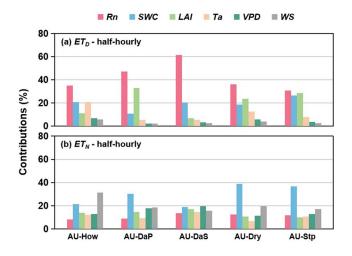




**Figure 6.** Relationships between daily vapor pressure deficit (VPD) and evapotranspiration (ET) under different soil water content (SWC) levels in the daytime (D; a) and at night (N; b) for the five study sites. The blue line was fitted with data of SWC > 24% (here, SWC = 24% was used for a demonstration purpose). The gray scatters were obtained by averaging every 500 points after sorting by VPD in a positive sequence.

that the hysteresis between ET and VPD has been suggested, which is regulated by biotic factors (e.g., leaf water potential) and abiotic factors (e.g., soil water status) (Zhang et al., 2014); however, this is beyond the scope of this study. In contrast, daily  $ET_N$  showed less response to nighttime VPD or even a negative pattern, which agreed well with previous studies (Barbour & Buckley, 2007; Resco De Dios et al., 2015). Since soil evaporation becomes more important in contributing to ET at night due to reduced plant Tr, the control of VPD on  $ET_N$  weakens (as shown here) in that VPD mainly affects ET through regulating stomatal water fluxes (Gu et al., 2006). The lower sensitivity of  $ET_N$  to VPD in our study was partly inconsistent with some previous findings, which might be due to different physiological strategies among different plant species (e.g., isohydric and anisohydric behaviors) (Ogle et al., 2012; O'keefe & Nippert, 2018). Moreover, when

SWC was no longer a limiting factor for ET (SWC > 24%), both  $ET_D$  and  $ET_N$  were positively correlated with VPD, attesting the interacting effect of VPD and SWC on ET processes.



**Figure 7.** Relative contributions of net radiation (Rn), soil water content (SWC), leaf area index (LAI), air temperature ( $T_a$ ), vapor pressure deficit (VPD), and wind speed (WS) to half-hourly daytime and non-gap-filled half-hourly nighttime evapotranspiration ( $ET_D$ ,  $ET_N$ ) for the five study sites from the boosted regression tree (BRT) analysis.

# 3.4. Quantification of Environmental Variables for Controlling Half-Hourly $ET_D$ and $ET_N$ Dynamics

Figure 7 compares the relative contributions of different environmental factors to half-hourly  $ET_D$  and nongap-filled half-hourly  $ET_N$  variations at each site based on the BRT results. The BRT results showed good agreement between observed and modeled half-hourly  $ET_D$  and  $ET_N$  (Table S4), illustrating the viability of using the BRT method for predicting  $ET_D$  and  $ET_N$  in the study region. During daytime, Rn was the dominant control at half-hourly timescales at the study sites, which supported recent findings that energy played more important role than water in regulating ET at diurnal timescales around the globe (for example, Huang et al., 2019; Scott & Biederman, 2019). Note that the high contribution of Rn at AUDaS was also consistent with the results shown in Figure 5, as the access to deeper soil moisture by deep-rooted trees at this site would alleviate the constraint of water supplies for ET throughout the year, resulting in the increased contribution of Rn. In addition, LAI was also an important

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contributor at the grassland sites (i.e., AU-DaP and AU-Stp), which was consistent with the reports of Williams and Torn (2015) and Han et al. (2020).

By comparison, the contrasts in the relative contributions of different variables to  $ET_N$  were less than to  $ET_D$ , suggesting that  $ET_N$  had more complex interactions with the surrounding environment. Particularly, half-hourly  $ET_N$  dynamics were mainly controlled by SWC and WS across the sites. For WS, the aerodynamic resistance that is related to WS regulates the water vapor transfer rate as a physical mechanism, and the wind can also prevent the formation of an intense nighttime inversion above the canopy (Green et al., 1989; Malek, 1992). In particular, at AU-How that is located offshore, the humid climate tended to promote the tight coupling between wind and  $ET_N$  (Meinzer et al., 1995).

# 3.5. Ecohydrological Implications

Increasing observational and numerical modeling evidence has indicated the important role played by  $ET_N$  in modulating terrestrial land surface and ecohydrological processes. The results of this study showed that long-term  $\overline{ET_N}/\overline{ET_D}$  was averaged at 8.6% along the NATT climate gradient. As a nonnegligible component of terrestrial ET, this has important implications for water resources management, particularly in arid regions as  $\overline{ET_N}/\overline{ET_D}$  increased with climate dryness. For instance,  $\overline{ET_N}$  (81.9 mm/year) at AU-Dry was about 3.6 times higher than the long-term surface runoff (~23 mm/year) observed by Jolly (2002) in the Dry River area, where the AU-Dry site is located.

Additionally, the observational data showed that the responses of  $ET_D$  and  $ET_N$  were different to environmental drivers at the study sites. For instance, at mean annual timescales, both vegetation and climatic conditions (e.g.,  $\overline{P}$  and  $\overline{VPD}$ ) were closely related to  $\overline{ET_D}$  variations across the sites, while  $\overline{ET_N}$  only varied with  $\overline{T_a}$ . As ET is one of the main subjects in climate change studies (Fisher et al., 2017; Novick et al., 2016), the differences in the nonlinear responses of  $ET_D$  and  $ET_N$  to those environmental drivers would have different consequences for regional and global water balances. Therefore, to improve our understanding of future climate change and its impact on hydrological cycles,  $ET_D$  and  $ET_N$  should be investigated separately.

Finally, we demonstrated that dominant controls on  $ET_D$  and  $ET_N$  also varied with temporal scales. For instance,  $ET_N$  dynamics were dominated by SWC from half-hourly to monthly timescales, while the primary controls on  $ET_N$  switched to SWC or energy supplies at annual timescales and to  $\overline{T_a}$  at mean annual timescales. As the selection of input variables needs to consider multitemporal scale correlations of ET with environmental drivers for land surface and climate models (Baldocchi & Wilson, 2001; Ogée et al., 2003), our data further showed that those multitemporal scale correlations occurred in both  $ET_D$  and  $ET_N$ . Furthermore, as ET and gross primary productivity (GPP) are strongly correlated, the different controlling mechanisms of ET (both  $ET_D$  and  $ET_N$ ) across temporal scales would affect the relative importance of multiple environmental factors to GPP variations over a wide range of timescales (Baldocchi & Wilson, 2001; Domec et al., 2012). This is critical for understanding water and carbon cycles in terrestrial ecosystems, particularly in light of changing climatic conditions.

## 4. Conclusions

Long-term EC records at five FLUXNET sites along a climate gradient in Northern Australia were analyzed to compare the responses of  $ET_D$  and  $ET_N$  to environmental variables at different temporal scales. The main conclusions of this study are as follows:

- 1. At mean annual timescales, as  $\overline{P}$  largely determined vegetation distributions in the study region, both vegetation and climatic conditions were closely related to  $\overline{ET_D}$  variations across the sites. From annual to daily timescales, water supplies (SWC) and vegetation conditions (LAI) controlled  $ET_D$  dynamics with SWC being more important at drier sites; whereas, at half-hourly timescales, meteorological conditions (mostly Rn) regulated  $ET_D$
- 2. Compared to  $ET_D$ ,  $ET_N$  was shown to be less responsive to environmental variables. At mean annual timescales, the primary controls on  $\overline{ET_N}$  were  $\overline{T_a}$  and  $\overline{Rn}$ , suggesting that with lower atmospheric water demands at night, the constraint of water supplies on  $\overline{ET_N}$  weakened. At annual timescales, the

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acquired and shared by the FLUXNET

community, including OzFlux-TERN

data are provided by ECMWF and

processed by LSCE. The FLUXNET

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networks. The ERA-Interim reanalysis

- controlling factors of  $ET_N$  depended on ecosystem types, SWC for grass sites and meteorological variables (i.e., Rn, VPD, and  $T_a$ ) for savanna sites. From monthly to half-hourly timescales,  $ET_N$  dynamics were dominated by SWC
- 3. At site levels, monthly  $ET_D$  and  $ET_N$  showed better correlations with meteorological and vegetation variables than annual  $ET_D$  and  $ET_N$ , and the coupling of  $ET_D$  and  $ET_N$  was also stronger at monthly timescales, particularly under drier climatic conditions. In addition, the responses of daily  $ET_D$  and  $ET_N$  to SWC exhibited similar threshold behaviors with higher sensitivities of daily  $ET_D$  to changes in SWC; whereas, daily  $ET_D$  and  $ET_N$  responded differently to VPD, especially under low VPD levels.

The above results proved our hypothesis that  $ET_D$  and  $ET_N$  had different responses to environmental drivers across diverse climate regimes and ecosystems at varying temporal scales. Our study provides a critical evaluation for contrasting  $ET_D$  and  $ET_N$  interactions with constantly changing environments, which has important implications for ecosystem water balance and land surface processes modeling.

# **Data Availability Statement**

Eddy covariance data are available at https://fluxnet.org/data/fluxnet2015-dataset/, and LAI data are available at https://e4ftl01.cr.usgs.gov/MOTA/.

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