

1 **Marine plastics alter the organic matter composition of the air-sea boundary layer, with**
2 **influences on CO₂ exchange: a large-scale analysis method to explore future ocean scenarios**

3 **List of authors:**

4 Luisa Galgani^{1,2,3*}, Eleni Tzempelikou^{4a}, Ioanna Kalantzi^{4b}, Anastasia Tsiola^{4b}, Manolis
5 Tsapakis^{4b}, Paraskevi Pitta^{4b}, Chiara Esposito⁵, Anastasia Tsotskou⁶, Iordanis Magiopoulos^{4b},
6 Roberto Benavides³, Tobias Steinhoff³ and Steven A. Loiselle^{1,7}

7 *corresponding author luisa.galgani@unisi.it

8 **Affiliations**

9 ¹Environmental Spectroscopy Group, Department of Biotechnology, Chemistry and Pharmacy,
10 University of Siena, Italy

11 ²Harbor Branch Oceanographic Institute of Florida Atlantic University, USA

12 ³GEOMAR-Helmholtz Center for Ocean Research Kiel, Germany

13 ^{4a}Institute of Oceanography, Athens, ^{4b}Institute of Oceanography, Heraklion, ^{2c}Institute of Marine
14 Biology Biotechnology & Aquaculture, Heraklion - Hellenic Centre for Marine Research, Greece.

15 ⁵Department of Bioscience - Lake Ecology and WATEC, Center for Water Technology, Aarhus
16 University, Denmark

17 ⁶University of Western Macedonia, School of Agricultural Sciences, Department of Agriculture,
18 Florina, Greece

19 ⁷Consorzio Interuniversitario Nazionale per la Scienza e Tecnologia dei Materiali, Florence, Italy.

20 **Keywords**

21 Microplastics, *p*CO₂, pH, sea-surface microlayer, mesocosms, marine gel particles, dissolved and
22 particulate organic matter

23 **Abstract**

24 Microplastics are substrates for microbial activity and can influence biomass production. This has
25 potentially important implications at the sea-surface microlayer, the marine boundary layer that
26 controls gas exchange with the atmosphere and where biologically produced organic compounds
27 can accumulate. In the present study, we used large scale mesocosms (filled with 3 m³ of seawater)
28 to simulate future ocean scenarios. We explored microbial organic matter dynamics in the sea-
29 surface microlayer in the presence and absence of microplastic contamination of the underlying
30 water. Our study shows that microplastics increased both biomass production and enrichment of
31 particulate carbohydrates and proteins in the sea-surface microlayer. Importantly, this resulted in
32 a 3% reduction in the concentration of dissolved CO₂ in the underlying water. This reduction
33 suggests direct and indirect impacts of microplastic pollution on the marine uptake of CO₂, by
34 modifying the biogenic composition of the sea's boundary layer with the atmosphere.

35 **Introduction**

36 The transition layer between environments is home to many fundamental physical, chemical and
37 biological processes. The sea-surface microlayer (SML) is a millimeter-sized interface between
38 the ocean and the atmosphere (Liss & Duce, 2005). It plays an essential role in ocean-climate
39 feedback by mediating air-sea gas exchange and marine aerosol emission (Wurl et al., 2017). The
40 SML has distinctly different biogeochemical properties to its underlying seawater and it is enriched
41 in organic matter, both dissolved organic matter (DOM) and particulate organic matter (POM), in
42 particular carbohydrate- and protein-rich marine gel particles (Engel et al., 2017; Liss & Duce,
43 2005).

44 The two major classes of marine gel particles present in the SML are Transparent Exopolymer
45 Particles (TEP) and Coomassie Stainable Particles (CSP). TEP and CSP give the SML its gel-like

46 composition which is prevalent in most parts of the ocean (Galgani et al., 2016; Wurl & Holmes,
47 2008) and support large and diversified microbial communities (Cunliffe & Murrell, 2009;
48 Cunliffe et al., 2011), sensitive to local environmental and meteorological conditions (Rahlff et
49 al., 2017; Zäncker et al., 2018). The enrichment of gel particles and their accompanying microbial
50 life favors the creation of stable surface films that can influence the air-sea fluxes of oxygen and
51 carbon dioxide (Calleja et al., 2013; Rahlff et al., 2019; Wurl et al., 2016).

52 Marine gel particles are derived from extracellular polymeric substances (EPS) released during
53 microbial metabolic functions. EPS exist in a continuum of sizes, including colloidal as well as
54 dissolved and particulate fractions (Decho & Gutierrez, 2017; Verdugo, 2012). TEP and CSP are
55 considered a class of EPS (Decho & Gutierrez, 2017), larger than 0.4 μm in relation to the pore
56 size of the filter used for their analysis (Engel, 2009).

57 The microbial release of exopolymers is enhanced on plastic surfaces (Michels et al., 2018), part
58 of the carbon-rich substrates that make up the resulting biofilm (Lear et al., 2021; Zhao et al.,
59 2021). The presence of microplastics in natural and artificial seawater can stimulate the microbial
60 release of DOM, probably due to a higher substrate availability for microbial growth (Boldrini et
61 al., 2021; Galgani et al., 2018). Likewise, both nanoparticles and microparticles can induce EPS
62 secretion by phytoplankton (Santschi et al., 2021; Shiu et al., 2020). The increased production of
63 organic matter around plastic particles can promote biogenic aggregates formation (Shiu et al.,
64 2020). These aggregates can move to the deep ocean (Galgani & Loiselle, 2021; Michels et al.,
65 2018) or remain in the SML (Galgani & Loiselle, 2019).

66 As plastic particles sustain niches for high microbial activity (Amaral-Zettler et al., 2020; Zettler
67 et al., 2013), one central hypothesis of the present study is that higher concentration of
68 microplastics in near surface conditions (< 3 m depths) stimulates a higher microbial production

69 of organic matter, adding on to the pool of organic compounds enriching the SML and thereby
70 modify the air-sea gas exchange properties of this interface.

71 It has been recently shown that seawater exposed plastic debris directly release climate relevant
72 gases like DMS (Savoca et al., 2016), methane and ethylene (Royer et al., 2018), suggesting that
73 high concentrations of surface plastics may have a direct effect on water-air interactions. It has
74 also been modeled that high concentrations of plastic can reduce the grazing pressure on
75 phytoplankton in marine regions where nutrients are not a limiting factor, with subsequent anoxic
76 conditions due to a cascade effect of initial high biomass production and degradation of the organic
77 material (Kvale et al., 2021). Since much of the plastic at sea is concentrated in oligotrophic
78 subtropical gyres (North Pacific, North Atlantic), it is expected that this additional carbon biomass
79 on plastics offers an additional carbon source able to alter biogeochemical cycles (Zhao et al.,
80 2021). Another central hypothesis of the present study was that higher microplastic concentrations
81 would promote high biomass production in oligotrophic conditions.

82 To test both hypotheses, pseudo-marine conditions needed to be created, with well-defined and
83 repeatable microplastic concentrations, but large enough to allow for aggregation, and for SML
84 organic aggregates interaction with bulk water. Likewise, the experimental conditions should
85 remain relatively stable over the medium term to allow for sampling of individual water masses
86 with no mixing of waters with different microplastic concentrations. This was achieved by six
87 large scale mesocosms, each filled with 3 m³ of oligotrophic seawater from the Sea of Crete. Three
88 of the mesocosms were amended with 30-µm diameter polystyrene microbeads (430 particles/L)
89 and three were plastic-free control mesocosms (< 0.5 particles larger than 1 micron/L). The SML
90 and bulk underlying water properties were compared over twelve days after the initial conditions,
91 prior to microplastic addition, were controlled for uniformity.

92 **Materials and methods**

93 *Mesocosms set up and SML sampling:* Six mesocosms with a height of ~2.5 m and a diameter of
94 1.32 m made of transparent polyethylene were gravity-filled with 3-m³ of coastal seawater pumped
95 from below the surface (2 m) in the bay of Gournes (Sea of Crete). To ensure homogeneity of
96 initial conditions, the water was divided equally in each of the six mesocosm and left overnight.
97 The density of seawater during the experiment was $1.032 \pm 0.001 \text{ g cm}^{-3}$ with an average salinity
98 of $41.4 \pm 1.6 \text{ PSU}$. For the duration of the experiment (12 days), the mesocosms were kept in a
99 150 m³ deep concrete tank with circulating seawater maintained at a constant temperature of 20 ± 1
100 °C. Each mesocosm was protected by a clear PVC lid to avoid atmospheric contamination.-The
101 first sampling (day 0) occurred the day after the mesocosms were filled. An aqueous solution of
102 30 µm diameter transparent polystyrene microbeads (Sigma-Aldrich, nr. 84135) with a density of
103 1.05 g cm^{-3} was added to three mesocosms (MP1-3) after the first sampling (day 1) for a
104 concentration of 430 microplastic L⁻¹, corresponding to about 5.92 µg C L^{-1} . Polystyrene is a very
105 abundant polymer found in oligotrophic marine areas (Pabortsava & Lampitt, 2020). Polystyrene
106 beads and polyethylene walls have a negative surface charge at pH above 2.5 (Beneš & Paulenová,
107 1973), reducing the possibility of attraction between the two materials in experimental conditions.
108 We choose not to clean the mesocosms walls from any possible biofilm formation: we believe that
109 this external procedure could significantly interfere with the parameters measured and be more
110 invasive than the effect of periphyton biofilm formed in a few days, which experimental studies
111 show being negligible for larger radius mesocosm (Chung-Chi & Kemp, 2004). Recent
112 experiments show biofilm formation on PE films after extended incubation times (> 60 and > 90
113 days) (Gupta & Devi, 2020; Han et al., 2020), not comparable to the present experiment.

114 The size range and concentration of polystyrene particles did not interfere with spectrophotometric
115 measurements of dissolved organic matter (Galgani et al., 2018; Galgani & Loiselle, 2019). While
116 a recent study has reported that virgin laboratory grade polymer and commercially available
117 polystyrene leaches dissolved organic carbon (DOC) in natural freshwater when exposed to dark
118 and light conditions (Lee et al., 2020), quantifiable DOC leaching was achieved at highly elevated
119 concentrations of 5 g L^{-1} , 10^6 times higher than the concentration used in the present study (6×10^{-6}
120 g L^{-1}).

121 This microplastic size also allows for comparison to other studies on microplastic influence on
122 marine biological processes like zooplankton ingestion (Cole et al., 2013; Cole et al., 2016) and
123 human health (Hwang et al., 2020).

124 Each mesocosm was continuously gently mixed through a centralized airlift system situated just
125 above the bottom surface to create a homogeneous distribution of the water, as described by Pitta
126 and colleagues (2016). The mixing system was the same in each mesocosm. Mesocosm set up,
127 manipulation, and sampling of the underlying water were conducted daily according to standard
128 methods for mesocosms studies performed at CretaCosmos facility
129 (<https://www.aquacosm.eu/mesocosm/cretacosmos/>) (Pitta et al., 2016; Tsiola, Tsagaraki, et al.,
130 2017). The sea-surface microlayer (SML) was sampled from each mesocosm early in the morning,
131 and prior to bulk water sampling on day 0, day 1, day 3, day 5, day 7, day 9 and day 10, to allow
132 a proper re-establishment of the surface film and minimized disturbance. The SML was sampled
133 contemporarily in each mesocosm with 30 cm x 30 cm silicate glass plates with an effective
134 sampling area of 1800 cm^2 . Glass plates were inserted into the mesocosms perpendicular to the
135 surface and withdrawn at a controlled rate of $\sim 6 \text{ cm s}^{-1}$ as suggested by Carlson (Carlson, 1982).
136 The glass plate approach collects a thinner SML ($\sim 60\text{--}150 \text{ }\mu\text{m}$) when compared to, e.g., the

137 Garrett screen (150–300µm) (Garrett, 1965). The glass plate method allowed sampling of
138 sufficient volume for analysis with a minimal dilution of the underlying water. The sample retained
139 on both sides of the plates was removed with a wiper and poured into bottles with the aid of a
140 funnel. The procedure was repeated until the necessary volume for analysis was obtained, tracking
141 the exact amount of dips per mesocosm. The first sample was discarded and used to rinse the
142 collecting bottle. Glass plates, collecting bottles, wipers and funnels were acid cleaned (HCl 10
143 %) and Milli-Q rinsed prior use. To avoid cross-contamination, the control mesocosms and the
144 microplastic-treated mesocosms had different sampling equipment (glass plates, funnels,
145 collection bottles, wipers).

146 The thickness (d , µm) of the sampled SML was estimated as:

$$147 \quad d = V / (A \times n) \quad (1)$$

148 where V is the SML volume collected, i.e., 60–140 mL, A is the sampling area of the glass plate
149 ($A = 1800 \text{ cm}^2$) and n is the number of dips. The apparent thickness of the SML ranged between
150 37 and 72 µm, with an overall mean of 54.7 ± 9.0 µm in agreement to previous studies (Carlson,
151 1982; Engel et al., 2018; Galgani & Engel, 2016; Rahlff et al., 2019; Zäncker et al., 2017). The
152 sampling thickness was similar for all mesocosms and sampling days. The samples were
153 immediately processed in the laboratory within maximum 30 minutes after collection.

154 ***Dissolved Inorganic Carbon (DIC) and Total Alkalinity (TA) measurements:*** Following Dickson
155 et al. (2007), seawater was drawn into 500 mL glass bottles using a tube to fill them from bottom
156 to the top. Approximately half of the bottle volume was overflowed and a small head-space
157 (approximately 1% of the bottle volume) was left to allow for water expansion. The samples were
158 fixed by adding 100 µL saturated mercuric chloride (HgCl_2) solution thus preventing further
159 biological activity and stored in the dark at room temperature until analysis. The DIC

160 measurements were performed using a coulometric technique with a SOMMA system (Johnson et
161 al., 1998). The determination of TA was performed by potentiometric titration using a VINDTA
162 system (Mintrop et al., 2000). Certified seawater samples were routinely measured to determine a
163 precision estimated to $2.8\mu\text{mol kg}^{-1}$ for DIC and $1.8\mu\text{mol kg}^{-1}$ for TA. ***Dissolved carbon dioxide***
164 ***estimates:*** The partial pressure of carbon dioxide in the mesocosms ($p\text{CO}_2$) as well as pH, CO_3^{2-} ,
165 HCO_3^- and CO_2 were calculated from DIC and TA at ambient temperature of $20\pm 1^\circ\text{C}$ and at sea-
166 level pressure with the CO2SYS program (Pierrot et al., 2006) ([https://cdiac.ess-
168 dive.lbl.gov/ftp/co2sys/](https://cdiac.ess-
167 dive.lbl.gov/ftp/co2sys/)). We applied the equilibrium constants K1 and K2 of Lueker et al. (2000)
169 as suggested by Dickson et al. (2007) and Orr et al. (2015) for a wider range of salinity and used
170 a K_{SO_4} value, the dissociation constant for HSO_4^- , of Dickson (1990).

170 ***Dissolved organic carbon (DOC) and chromophoric dissolved organic matter (CDOM):*** Samples
171 for DOC were filtered immediately after sampling in duplicate through $0.2\mu\text{m}$ polycarbonate
172 membranes and stored cool ($+4^\circ\text{C}$) in pre-combusted glass ampoules until analysis. Concentrations
173 were determined using a Shimadzu TOC-V organic carbon analyser and following the high
174 temperature catalytic oxidation method. The system was standardized prior to analysis using a
175 potassium hydrogen phthalate standard solution. Each sample was injected 3 to 5 times and DOC
176 concentrations were calculated from the average value of three replicates that yielded a relative
177 standard deviation $<2\%$. Analytical precision and accuracy were tested against Deep Atlantic
178 Seawater Reference Material provided by the DOC-CRM program (University of Miami – D.A.
179 Hansell, batch 16); measured values: 0.510-0.580 ($n=10$), certified value: 0.516-0.540. Drift
180 correction of the DOC results was applied as needed.

181 Samples for CDOM were filtered immediately after collection in duplicates through $0.2\mu\text{m}$
182 polyethersulfone syringe filters and stored cool ($+4^\circ\text{C}$) in pre-combusted amber glass vials until

183 analysis, which was performed within 4 weeks. The CDOM absorbance spectrum was measured
184 with a Lambda 10 ultraviolet-visible light (UV-Vis) Spectrophotometer (Perkin Elmer) from 210
185 to 750 nm at 960 nm/min, 1 nm wavelength resolution, and at room temperature ($20^{\circ}\text{C} \pm 2^{\circ}\text{C}$). For
186 each analysis, spectra were corrected for baseline, Milli-Q water absorbance and for scattering by
187 subtracting the absorbance values at 730 nm. Absorption coefficients $a(\lambda)$ were calculated from
188 absorbance values after Bricaud et al. (Bricaud et al., 1981). The CDOM absorption spectral slope
189 S (nm^{-1}) was determined by linear regression of log-transformed absorption spectra against the
190 wavelength (Bricaud et al., 1981):

$$191 \quad a(\lambda) = a_0 e^{-S(\lambda-\lambda_0)} \quad (2)$$

192 With $a(\lambda_0)$ being the absorption coefficient at a reference wavelength λ_0 . We used multiple 20-nm
193 wavelength intervals in a stepwise (1 nm) linear regression analysis according to Loiselle et al.
194 (Loiselle et al., 2009). Spectral slope correlates with changes in CDOM due to irradiation
195 (photobleaching), and in the wavelength range 275-295, $S_{(275-295)}$ has been shown to be inversely
196 related with DOM molecular weight (Helms et al., 2008).

197 ***Autotrophic and heterotrophic microorganisms:*** Abundances of autotrophic and heterotrophic
198 microorganisms were measured by flow cytometry. Samples for heterotrophic bacteria were fixed
199 with 25% 0.2 μm -filtered glutaraldehyde (0.5% final concentration), incubated at 4°C for 45 min,
200 flash frozen in liquid nitrogen and stored at -80°C until analysis. Frozen samples were thawed at
201 room temperature and sub-samples were stained for bacterial enumeration with the nucleic-acid
202 stain SYBR Green I (final dilution 4×10^{-4} of the stock solution in Tris-EDTA buffer, $\text{pH} = 8$)
203 and incubated for 10 min in the dark (Marie et al., 1997). Sample fluorescence signal to distinguish
204 high and low DNA content cells. Samples for autotrophic microorganisms were not fixed and were

205 analysed without prior staining, based on their auto-fluorescence signals. A FACSCalibur™ flow
206 cytometer (Becton Dickinson) was used following Tsiola et al. (2017).

207 **Marine gels:** Total area and numbers of gel particles were determined by optical microscopy
208 (Engel, 2009). Ten to fifteen mL of sample were filtered using 0.2 µm Nuclepore membranes
209 (Whatman) and stained with 1 mL Alcian Blue solution for polysaccharidic gels (Transparent
210 Exopolymer Particles, TEP) and 1 mL Coomassie Brilliant Blue G solution for proteinaceous gels
211 (Coomassie Stainable Particles, CSP). Filters were mounted onto Cytoclear® slides and stored at
212 – 20°C until microscopic analysis. For each slide, thirty images were taken randomly at 200x
213 magnification with a light microscope equipped with a digital camera. The analysis of the cross-
214 sectional area of marine gels was performed with an image analysis software (ImageJ, U.S.
215 National Institutes of Health) and used to calculate the equivalent spherical diameter (ESD) of
216 individual particles, particle number, volume and total area. The size frequency distribution of
217 marine gel particles was determined according to their equivalent spherical diameter, described
218 with a power function of the type:

$$219 \frac{dN}{d(d_p)} = k_p^\delta \quad (3)$$

220 with dN as the number of particles per unit volume in the size range d_p to $[d_p + d(d_p)]$, k a constant
221 which depends on the concentrations of particles, and δ the slope ($\delta < 0$) describing the size
222 distribution. A less negative δ implies an increase in the fraction of larger marine gels. k and δ
223 were both derived from regressions of $\log[dN/d(d_p)]$ versus $\log[d_p]$ (Harlay et al., 2009; Mari &
224 Burd, 1998; Mari & Kiørboe, 1996). The volume concentration of TEP and CSP refers to the mean
225 volume of the particles $> 0.2 \mu\text{m}$ (membrane pore size cutoff); changes in this parameter indicate
226 modifications in particle dynamics such as aggregation/disaggregation processes.

227 Since TEP are considered fractal aggregates, the volume and the carbon content of these marine
 228 gel particles are assumed to be proportional to r^D , with r being the equivalent spherical radius (μm)
 229 and D the fractal scaling dimension associated with the size-distribution of marine gels (Engel,
 230 2009; Mari & Burd, 1998; Mari & Kiørboe, 1996). TEP carbon content (TEP-C, expressed in μg
 231 L^{-1}) was determined from marine gel size spectra according to Mari (Mari, 1999) and Engel (2009):
 232
$$\text{TEP-C } [\mu\text{g L}^{-1}] = 0.25 \times 10^{-6} r^D \quad (4)$$

233 With $D = - 2.55$

234 **Enrichment Factors:** To determine the enrichment or the depletion of each parameter analysed in
 235 the SML compared to the underlying water, we determined the Enrichment Factor (EF), defined
 236 as:

$$237 \text{ EF} = ([X]_{\mu} / [X]_{\text{b}}) \quad (5)$$

238 With $[X]_{\mu}$, $[X]_{\text{b}}$ the concentration of the specific parameter in the SML (μ) or underlying water (b)
 239 (Liss & Duce, 2005). An $\text{EF} = 1$ indicates that SML and underlying water values are similar, thus
 240 no enrichment or depletion in the SML. While bubbling may have promoted the enrichment of
 241 certain compounds in the SML, this process occurred in all mesocosms thus the influence of the
 242 airlift mixing system in the comparison of the dynamics of plastic free versus plastic enriched
 243 treatments is negligible.

244 **Data analysis and statistics:** To highlight the treatment effect (microplastic addition) and avoid
 245 the temporal variability, we calculated the normalized anomaly y_{ij} of each mesocosm (j) per day
 246 ($i = 0, \dots, 11$) from the overall daily mean of the mesocosms $\bar{y}_i = \frac{1}{6} \sum_j^6 (x_j)_i$ following a
 247 procedure often applied in mesocosms studies (Endres et al., 2014; Engel et al., 2013):

$$248 y_{ij} = (x_{ij} - \bar{y}_i) / \bar{y}_i \quad (6)$$

249 Differences between control and treated mesocosms were determined by two-tailed unpaired t-
250 tests and Mann-Whitney tests on normalized anomalies, depending on the distribution of the data.
251 Repeated Measures Two-Way ANOVA and Mixed Effects Model (REML) were also used to
252 analyse temporal variations between the control and microplastic treated mesocosms, the latter
253 used in case of missing observations. This method is widely used to analyse data from mesocosms
254 experiments (Dimitriou et al., 2017; Rahav et al., 2016). The fixed factor considered is the
255 treatment (microplastics addition / control) and the random effect is time (days). Correlations
256 among parameters were determined by Multiple Linear Regression and Spearman correlation
257 analysis. Statistical significance was accepted for $p < 0.05$ and considering a Bonferroni correction
258 for multiple comparisons. All statistical tests were performed with Prism 8.02 (GraphPad
259 Software, San Diego, CA, USA) and Minitab18 (Minitab Inc., USA).

260

261 **Results**

262 An earlier study reported an increase in the production of POM and marine gels in the bulk waters
263 of the plastic-amended mesocosms with respect to the plastic-free controls (Galgani et al., 2019).
264 In the present study, we explore differences in the composition of the SML and the possible
265 influence of different SML compositions on the $p\text{CO}_2$ of the underlying bulk water. The SML was
266 analysed for marine gel particles (TEP and CSP), autotrophic and heterotrophic microbial
267 organisms, and dissolved organic matter parameters: Dissolved Organic Carbon, DOC,
268 Chromophoric Dissolved Organic Matter, CDOM, and spectral slope. Salinity, total alkalinity and
269 dissolved inorganic carbon (DIC) were measured from the underlying water and *in-situ* $p\text{CO}_2$, pH,
270 CO_3^{2-} , HCO_3^- and CO_2 were retrieved accordingly. In the following paragraphs, all parameters
271 described refer to the SML, except where specified. Biogeochemical processes and relevant data

272 of the underlying water are described in a previous publication (Galgani et al., 2019) and were not
273 the focus of this study.

274 *Sea surface microlayer dynamics and underlying water pCO₂*

275 Microplastic-amended mesocosms (MP) had significantly lower values of pCO₂ in the underlying
276 water and higher pH compared to control mesocosms (Figure 1). Mean values of pCO₂ in the MP
277 treatments were 3% lower than those found in control mesocosms, corresponding to an increase
278 of 0.14% in pH units (Table 1). Total alkalinity (TA) and dissolved inorganic carbon (DIC) were
279 also significantly higher in the underlying water of MP treatments (Figures 2 and S1). Similar
280 differences were found for estimated concentrations of CO₃²⁻ and CO₂, with higher CO₃²⁻ in MP
281 treatments (Figure 2).

282 Higher concentrations of both polysaccharidic TEP and proteinaceous CSP were measured in the
283 SML of the MP treated mesocosms (Figure 1, Table 1). An 30% increase in TEP and relative
284 carbon content (TEP-C) along with higher concentrations and particle abundances of both marine
285 gels occurred in the SML of the MP mesocosms, with a mean±SEM (standard error of means)
286 TEP-C of 424.8±58.5 µg C L⁻¹ compared to 324.3±42.9 µg C L⁻¹ in control mesocosms (Figure 1).
287 This was accompanied by a 1% increase in *Synechococcus*, and by a 23% increase of high-DNA
288 containing cells in the SML (Figures 1 and S2, Table 1), likely to be the main source of TEP for
289 the whole system. In the underlying water of the mesocosms, an initial (and rapidly declined)
290 phytoplankton bloom was attributed to the presence of autotrophic picoeukaryotes, while
291 *Synechococcus* growth showed a constant increase and represented the dominant species in the
292 mesocosms. This was expected as this species dominates in the oligotrophic waters of the Sea of
293 Crete at this time of the year (Galgani et al., 2019). Heterotrophic bacteria concentrations in the
294 SML were similar in MP and control mesocosms, and significant differences were observed over

295 time rather than across treatments (Figure S2, Table 1). Their concentrations were negatively
296 correlated to those of *Synechococcus*, indicating a potential competition within the highly dynamic
297 environment of the SML (Spearman $r = -0.40$, $p = 0.015$, $n = 37$). Alternating influence of primary
298 and secondary production in the SML was also evidenced in the dynamics of the spectral slope
299 $S_{275-295}$ of the chromophoric dissolved organic matter (CDOM) pool. Higher $S_{275-295}$ is associated
300 to lower molecular weight CDOM often resulting from degradation processes, while lower $S_{275-295}$
301 characterizes the production of higher molecular weight CDOM (Helms et al., 2008).

302 While DOC and CDOM measured as the absorption coefficient at 355 nm were similar between
303 control and MP mesocosms and significant variations were observed over time (Table 1, Figure
304 S3), $S_{275-295}$ instead was higher in the SML of MP mesocosms (Table 1, Figure S3) confirming that
305 there were significant differences in the DOM production and degradation processes between
306 treatments. In the SML, we observed a strong negative correlation between *Synechococcus* and
307 $S_{275-295}$ (Table 2), and a weaker positive correlation between heterotrophic bacteria and $S_{275-295}$.
308 $S_{275-295}$ was positively correlated to CSP ($\text{mm}^2 \text{L}^{-1}$), DOC and $p\text{CO}_2$ (Table 2) suggesting different
309 production dynamics of CSP with respect to TEP and linking CSP concentration to organic matter
310 degradation. In the SML, the amount of TEP-C was weakly related to the abundance of
311 *Synechococcus* (Linear Regression $R^2 = 0.12$, $p = 0.03$) suggesting a local source (i.e., within the
312 SML) of fresh biological production for this class of marine gel compounds. A positive correlation
313 between DOC and $S_{275-295}$ suggested that the pool of DOC in the SML of the mesocosms is mostly
314 composed of degraded or reworked material, and unlikely linked to a fresh biomass production.

315 Most importantly, we observed significant correlations between the $p\text{CO}_2$ in the underlying water
316 (as well as pH, TA, DIC, CO_3^{2-} and CO_2) with the concentration of TEP-C and *Synechococcus*
317 cells in the SML (Table 3). $p\text{CO}_2$ decreased with increasing TEP concentration and carbon content

318 (Figure S4a). Instead, CSP in the SML had the opposite relationship with $p\text{CO}_2$ with respect to
319 TEP (Figure 3). This was further demonstrated in the linear relationship between $p\text{CO}_2$ and gel
320 concentrations (Equation 7):

$$321 \quad p\text{CO}_2 [\mu\text{atm}] = 413.189 - 0.0509 \text{ TEP} [\text{mm}^2 \text{ L}^{-1}] + 0.04708 \text{ CSP} [\text{mm}^2 \text{ L}^{-1}] \quad (7)$$

$$322 \quad R^2 = 0.38, \quad p < 0.001, \quad F=11.98$$

323 It should be noted that underlying water TEP and CSP concentrations from the same experiment
324 did not show any relation to $p\text{CO}_2$. Together with SML gel concentrations, $p\text{CO}_2$ in the underlying
325 water was related to SML *S₂₇₅₋₂₉₅* and SML *Synechococcus* abundance (Figures S4b, S4c and
326 Tables 2 and 3), indicating the relationship of $p\text{CO}_2$ to different phases of organic matter cycling.

327 *Sea surface microlayer enrichment and relation to underlying water parameters*

328 In this experiment, the SML represented an enriched environment with respect to the bulk water
329 conditions (Figure 4). *Synechococcus* in the SML was strongly related to the underlying water cell
330 abundances (Spearman $r = 0.945$, $p < 0.0001$, $n = 39$), as well as the amount of CDOM (Spearman
331 $r = 0.859$, $p < 0.0001$, $n = 42$), the latter pointing to similar dynamics in DOM turnover between
332 SML and underlying water. TEP concentration ($\text{mm}^2 \text{ L}^{-1}$) in the SML also showed a low but
333 significant relation to underlying water TEP to some extent (Spearman $r = 0.359$, $p = 0.0197$, n
334 $= 42$), confirming that most of the TEP variability in the SML was associated to *Synechococcus*
335 cell abundances.

336 Particularly high enrichment factors were observed for marine gels in the SML in all mesocosms,
337 as expected. However, the enrichment factors for TEP and CSP were significantly higher in the
338 MP mesocosms with respect to plastic-free controls (Figure 5, Table S1) indicating an increased
339 accumulation of these gels in the SML of the mesocosms where plastic was added.

340 It should be noted that the concentration of plastic particles used in this study is higher than typical
341 marine conditions and more similar to those of impacted coastal areas and projections of future
342 conditions. Plastic production is expected to double in the next decades and in the absence of
343 efficient management practices, will likely increase the amount of plastic in our oceans (Geyer et
344 al., 2017). High concentrations of microplastics (10^5 particles m^{-3}) have been already found in a
345 Swedish harbour in proximity of a plastic manufacture plant (Lindgren et al., 2016).

346 **Discussion**

347 The twelve-day experiment coincided with a *Synechococcus* bloom phase in all six mesocosms,
348 during which autotrophic production of particulate organic matter prevailed over heterotrophic
349 production. In the MP mesocosms there was an increased abundance of *Synechococcus* in the
350 underlying water, which dominated over autotrophic picoeukaryotes cell numbers by an order of
351 magnitude (Galgani et al., 2019). Microbial attachment on particles is a general phenomenon in
352 aquatic ecosystems (Paerl, 1975) and microplastics may serve as a physical support for different
353 microorganisms (Zettler et al., 2013; Zhao et al., 2021) creating hot spots of high metabolic activity
354 (Dang & Lovell, 2015).

355 *Synechococcus* is an important primary producer in the Sea of Crete and a known TEP producer
356 also in nutrient-limiting conditions (Deng et al., 2016; Ortega-Retuerta et al., 2019). Globally,
357 *Synechococcus* contributes to 21% of total CO₂ fixation (Jardillier et al., 2010).

358 We hypothesize that two concurrent processes led to the increased accumulation of TEP in the
359 surface microlayer of the mesocosms containing MPs. Firstly, MP related increases in
360 *Synechococcus* production led to increased TEP accumulation in the underlying water and
361 enrichment in the SML. The relationship between underlying water and SML gel particles is not
362 novel in field (Engel & Galgani, 2016) and mesocosms studies (Galgani et al., 2014). Secondly,

363 *Synechococcus* cells can produce TEP directly in the SML (Yue et al., 2018) and their increased
364 concentration in the SML in the mesocosms containing MPs was an additional source for TEP
365 accumulation.

366 In the ocean, TEP aggregation is proved to be a sink for marine carbon (Engel et al., 2004) and
367 TEP production as a response to increased CO₂ has been demonstrated for natural plankton
368 communities (Engel, 2002). It has also been shown that microbial activity in the sea-surface
369 microlayer can control the exchange rate of atmospheric CO₂ (Calleja et al., 2005), and in our
370 experiment, TEP accumulation in the SML of the MP mesocosms may have served as a sink for
371 CO₂ as described in equation (7).

372 To further support this difference, DIC partitioning in the underlying water of MP mesocosms had
373 higher CO₃²⁻ and lower CO₂ concentrations.

374 *Synechococcus* has also been associated to “whiting” events in marine and lakes environments
375 (Dittrich et al., 2003; Thompson et al., 1997). Whiting events occur when carbonates of biogenic
376 origin are precipitated by microorganisms like cyanobacteria through photo-and chemosynthetic
377 autotrophy in the presence of Mg and Ca counterions (Dittrich et al., 2003; Thompson, 2000;
378 Thompson et al., 1997). Many bacteria species are implicated in CaCO₃ precipitation, which
379 represents a potential mechanism for CO₂ sequestration and generally occurs at high pH, during
380 active photosynthesis, and when DIC is limiting, such as after a bloom (Callieri, 2017). During the
381 twelve-day experiment, the highest abundance of *Synechococcus* in the SML and underlying water
382 appeared after the picoeukaryote and heterotrophic blooms (Galgani et al., 2019), corresponding
383 to a migration and enrichment of organic compounds to the SML.

384 In particular, the enrichment of organic particles (i.e. TEP and CSP) in the SML of MP mesocosms
385 led to the establishment of a highly enriched surface film, potentially able to modify gas exchange

386 between the mesocosms and the surrounding atmosphere. The different partitioning of DIC
387 between CO_3^{2-} and CO_2 in MP treatments confirms an increased autotrophy in MP mesocosms that
388 led to lower $p\text{CO}_2$ beneath. In control mesocosms, heterotrophy did not increase, rather,
389 autotrophic production was less as additional microplastic substrates for growth and metabolism
390 were missing. This led to an overall lower production of marine gels in the underlying water and
391 in the SML, resulting in a less enriched surface film and an increased exchange of atmospheric
392 CO_2 at the air-sea interface. This is similar to the presence of surfactants in the sea-surface
393 microlayer, where the resulting laminar diffusion layer reduces gas transfer (Frew, 1997). Some
394 biogenic surfactants and surface slicks carrying a high microbial biomass can reduce air-sea CO_2
395 exchange by 15% to 19% (Mustaffa et al., 2020; Wurl et al., 2016).

396 While initially $p\text{CO}_2$ concentration might have been lower in the MP mesocosms, we should note
397 that SML enrichment does not only limit the exchange of atmospheric CO_2 , but also that of oxygen
398 across the air-sea interface. In the post-bloom phase dominated by the heterotrophic
399 remineralization of organic matter, CO_2 is put back into the system through microbial respiration,
400 while oxygen concentrations are reduced. This may be especially important in coastal, shallow and
401 semi-enclosed marine areas affected by plastic and other types of organic or nutrient pollution that
402 favour high autotrophic biomass production. In areas where zooplankton feed on plastic particles
403 because of high plastic concentration, grazing pressure on primary producers is reduced, further
404 increasing autotrophic biomass in eutrophic waters, where the subsequent remineralization of
405 organic matter can further reduce oxygen concentrations (Kvale et al., 2021).

406 Our study shows that microplastics can increase the accumulation of marine gel particles in the
407 SML by a 25-30% compared to plastic-free conditions. By using large scale mesocosms, it was
408 possible to explore daily changes in SML formation and the resulting impact on the underlying

409 water $p\text{CO}_2$. This supports a better understanding of localized anoxic or hypoxic zones often
410 observed in estuarine and upwelling areas where the SML plays an essential role in air-sea gas
411 exchange (Engel & Galgani, 2016; Hepach et al., 2016; Upstill-Goddard, 2006) and where marine
412 plastisphere communities can have direct effect on the concentration of N_2O and CO_2 (Cornejo-
413 D'Ottone et al., 2020; Su et al., 2022).

414 While this method demonstrated the indirect effect of plastics on seawater $p\text{CO}_2$ concentration
415 through the SML, another effect that should be considered is the direct impact on the production
416 and response of marine gels in the SML. TEP and CSP are distinct, insoluble macromolecules
417 derived from the aggregation and annealing of DOM polymeric precursors produced during
418 microbial growth and metabolism (Cisternas-Novoa et al., 2015; Engel, 2009; Thornton, 2018;
419 Thornton et al., 2016). TEP concentrations in the SML were only partly related to those of the
420 underlying water, indicating that an additional source of TEP may be the SML itself through the
421 microbial activity of *Synechococcus*. As such, the SML may act as a direct sink of atmospheric
422 CO_2 . CSP instead seemed to be more clearly related to the degradation of organic matter present
423 in the SML, as CDOM and $S_{275-295}$ measurements indicated their higher lability and rapid turnover
424 (Thornton, 2018). This creates a pool of organic matter in the SML that is completely independent
425 from underlying water processes, a phenomenon observed in highly productive marine regions
426 (Galgani & Engel, 2016; Zäncker et al., 2017).

427 The present mesocosm study shows an important new impact of plastic pollution on marine carbon
428 biogeochemistry that directly affects processes at the air-sea boundary layer. The increased
429 production and accumulation of these two different classes of marine gels in the SML represent an
430 important mechanism regulating air sea gas exchange, both indirectly through an enriched surface
431 film, as well as directly as the site for high remineralization of organic matter.

432 **Conclusions**

433 An increasing number of studies suggests that plastic pollution influences the marine biological
434 pump and the ocean's capability to store the increasing atmospheric amounts of anthropogenic
435 carbon dioxide (Galgani & Loiselle, 2021; Kvale et al., 2021; Shen et al., 2020). Our results show
436 the impact of plastic pollution at the very surface of the ocean, with direct and indirect mechanisms
437 that control the ocean CO₂ exchange with the atmosphere. Further efforts should be made to
438 explore these mechanisms across different marine conditions and autotrophic species. Clearly,
439 plastic pollution impacts on marine biogeochemistry go well beyond that explored to date, and
440 should be included in international agreements for their potential effects on the functionality of the
441 ocean carbon sink (Cooley et al., 2019).

442 **Acknowledgements**

443 We greatly acknowledge G. Piperakis for the setting up of the mesocosms and for his technical
444 assistance throughout the experiment. K. Mylona, I. Santi, S. Zivanovic, E. Dafnomili, S.
445 Diliberto, and A. Loiselle are greatly acknowledged for technical support.

446 **Author contributions:** **Author contributions:** LG designed the experiment in consultation with
447 SAL, MT and PP, analysed the data in consultation with SAL and wrote the manuscript. LG, ET,
448 IK, AT, MT, PP, CE, ATs, IM, and SAL contributed to set up and experiment running, samples
449 analysis and manuscript editing. RB and TS contributed to samples analysis and manuscript
450 editing.

451 **Funding:** This work received funding from the European Union's Horizon 2020 Research and
452 Innovation Programme under the Marie Skłodowska-Curie grant agreement No. 702747 –
453 *POSEIDOMM*, to L. Galgani.

454 **Competing Interests:** The authors declare no competing interests.

455 **Data availability:** all data will be made available on an open repository after publication in
456 conformity with the requirements of all Horizon 2020 funded research projects.

457 **References**

- 458 Amaral-Zettler, L. A., Zettler, E. R., & Mincer, T. J. (2020). Ecology of the plastisphere. *Nat. Rev.*
459 *Microbiol.*, 18(3), 139-151. <https://doi.org/10.1038/s41579-019-0308-0>
- 460 Beneš, P., & Paulenová, M. (1973). Surface charge and adsorption properties of polyethylene in aqueous
461 solutions of inorganic electrolytes. *Kolloid-Zeitschrift und Zeitschrift für Polymere*, 251(10), 766-
462 771. <https://doi.org/10.1007/BF01499104>
- 463 Boldrini, A., Galgani, L., Consumi, M., & Loiseau, S. A. (2021). Microplastics Contamination versus
464 Inorganic Particles: Effects on the Dynamics of Marine Dissolved Organic Matter. *Environments*,
465 8(3), 21. <https://doi.org/https://doi.org/10.3390/environments8030021>
- 466 Bricaud, A., Morel, A., & Prieur, L. (1981). Absorption by dissolved organic matter of the sea (yellow
467 substance) in the UV and visible domains. *Limnol. Oceanogr.*, 26, 43-53.
468 <https://doi.org/https://doi.org/10.4319/lo.1981.26.1.0043>
- 469 Calleja, M. L., Duarte, C. M., Álvarez, M., Vaquer-Sunyer, R., Agustí, S., & Herndl, G. J. (2013).
470 Prevalence of strong vertical CO₂ and O₂ variability in the top meters of the ocean. *Global*
471 *Biogeochem. Cycles*, 27(3), 941-949. <https://doi.org/10.1002/gbc.20081>
- 472 Calleja, M. L., Duarte, C. M., Navarro, N., & Agustí, S. (2005). Control of air-sea CO₂ disequilibria in the
473 subtropical NE Atlantic by planktonic metabolism under the ocean skin. *Geophysical Research*
474 *Letters*, 32(8). <https://doi.org/https://doi.org/10.1029/2004GL022120>
- 475 Callieri, C. (2017). Synechococcus plasticity under environmental changes. *FEMS Microbiol. Lett.*,
476 364(23). <https://doi.org/10.1093/femsle/fnx229>
- 477 Carlson, D. J. (1982). Phytoplankton in marine surface microlayers. *Can. J. Microbiol.*, 28(11), 1226-1234.
478 <https://doi.org/10.1139/m82-183>
- 479 Chung-Chi, C., & Kemp, W. M. (2004). Periphyton communities in experimental marine ecosystems:
480 scaling the effects of removal from container walls. *Marine Ecology Progress Series*, 271, 27-41.
481 <https://www.int-res.com/abstracts/meps/v271/p27-41/>
- 482 Cisternas-Novoa, C., Lee, C., & Engel, A. (2015). Transparent exopolymer particles (TEP) and Coomassie
483 stainable particles (CSP): Differences between their origin and vertical distributions in the ocean.
484 *Mar. Chem.*, 175, 56-71. <https://doi.org/https://doi.org/10.1016/j.marchem.2015.03.009>
- 485 Cole, M., Lindeque, P., Fileman, E., Halsband, C., Goodhead, R., Moger, J., & Galloway, T. S. (2013).
486 Microplastic Ingestion by Zooplankton. *Environ. Sci. Technol.*, 47(12), 6646-6655.
487 <https://doi.org/10.1021/es400663f>
- 488 Cole, M., Lindeque, P. K., Fileman, E., Clark, J., Lewis, C., Halsband, C., & Galloway, T. S. (2016).
489 Microplastics Alter the Properties and Sinking Rates of Zooplankton Faecal Pellets. *Environ. Sci.*
490 *Technol.*, 50(6), 3239-3246. <https://doi.org/10.1021/acs.est.5b05905>
- 491 Cooley, S. R., Bello, B., Bodansky, D., Mansell, A., Merkl, A., Purvis, N., Ruffo, S., Taraska, G., Zivian,
492 A., & Leonard, G. H. (2019). Overlooked ocean strategies to address climate change. *Glob.*
493 *Environ. Change*, 59, 101968. <https://doi.org/https://doi.org/10.1016/j.gloenvcha.2019.101968>

- 494 Cornejo-D'Ottone, M., Molina, V., Pavez, J., & Silva, N. (2020). Greenhouse gas cycling by the
495 plastisphere: The sleeper issue of plastic pollution. *Chemosphere*, 246, 125709.
496 <https://doi.org/https://doi.org/10.1016/j.chemosphere.2019.125709>
- 497 Cunliffe, M., & Murrell, J. C. (2009). Eukarya 18S rRNA gene diversity in the sea surface microlayer:
498 implications for the structure of the neustonic microbial loop. *ISME J.*, 4(3), 455-458.
499 <https://doi.org/10.1038/ismej.2009.133>
- 500 Cunliffe, M., Upstill-Goddard, R. C., & Murrell, J. C. (2011). Microbiology of aquatic surface microlayers.
501 *FEMS Microbiol. Rev.*, 35(2), 233-246. <https://doi.org/10.1111/j.1574-6976.2010.00246.x>
- 502 Dang, H., & Lovell, C. R. (2015). Microbial Surface Colonization and Biofilm Development in Marine
503 Environments. *Microbiol. Mol. Biol. Rev.: MMBR*, 80(1), 91-138.
504 <https://doi.org/10.1128/MMBR.00037-15>
- 505 Decho, A. W., & Gutierrez, T. (2017). Microbial Extracellular Polymeric Substances (EPSs) in Ocean
506 Systems. *Frontiers in Microbiology*, 8. <https://doi.org/10.3389/fmicb.2017.00922>
- 507 Deng, W., Cruz, B. N., & Neuer, S. (2016). Effects of nutrient limitation on cell growth, TEP production
508 and aggregate formation of marine *Synechococcus*. *Aquat. Microb. Ecol.*, 78(1), 39-49.
509 <https://www.int-res.com/abstracts/ame/v78/n1/p39-49/>
- 510 Dickson, A. G. (1990). Standard potential of the reaction: $\text{AgCl(s)} + 12\text{H}_2\text{(g)} = \text{Ag(s)} + \text{HCl(aq)}$, and and
511 the standard acidity constant of the ion HSO_4^- in synthetic sea water from 273.15 to 318.15 K. *J.*
512 *Chem. Thermodyn.*, 22(2), 113-127. [https://doi.org/https://doi.org/10.1016/0021-9614\(90\)90074-Z](https://doi.org/https://doi.org/10.1016/0021-9614(90)90074-Z)
- 513 Dickson, A. G., Sabine, C. L., & Christian, J. R. E. (2007). *Guide to Best Practices for Ocean CO2*
514 *Measurements*. PICES Special Publication 3, 191 pp. [https://www.ncei.noaa.gov/access/ocean-](https://www.ncei.noaa.gov/access/ocean-carbon-acidification-data-system/oceans/Handbook_2007.html)
515 [carbon-acidification-data-system/oceans/Handbook_2007.html](https://www.ncei.noaa.gov/access/ocean-carbon-acidification-data-system/oceans/Handbook_2007.html)
- 516 Dimitriou, P. D., Papageorgiou, N., Geropoulos, A., Kalogeropoulou, V., Moraitis, M., Santi, I.,
517 Tsikopoulou, I., Pitta, P., & Karakassis, I. (2017). A novel mesocosm setup for benthic-pelagic
518 coupling experiments. *Limnology and Oceanography: Methods*, 15(4), 349-362.
519 <https://doi.org/https://doi.org/10.1002/lom3.10163>
- 520 Dittrich, M., Müller, B., Mavrocordatos, D., & Wehrli, B. (2003). Induced Calcite Precipitation by
521 Cyanobacterium *Synechococcus*. *Acta Hydrochim. Hydrobiol.*, 31(2), 162-169.
522 <https://doi.org/https://doi.org/10.1002/aheh.200300486>
- 523 Endres, S., Galgani, L., Riebesell, U., Schulz, K.-G., & Engel, A. (2014). Stimulated Bacterial Growth
524 under Elevated pCO₂: Results from an Off-Shore Mesocosm Study. *PLOS ONE*, 9(6), e99228.
525 <https://doi.org/10.1371/journal.pone.0099228>
- 526 Engel, A. (2002). Direct relationship between CO₂ uptake and transparent exopolymer particles production
527 in natural phytoplankton. *J. Plankton Res.*, 24(1), 49-53. <https://doi.org/10.1093/plankt/24.1.49>
- 528 Engel, A. (2009). Determination of Marine Gel Particles. In O. E. Wurl (Ed.), *Practical Guidelines for the*
529 *Analysis of Seawater*. CRC Press. <https://doi.org/doi:10.1201/9781420073072.ch7>
- 530 Engel, A., Bange, H. W., Cunliffe, M., Burrows, S. M., Friedrichs, G., Galgani, L., Herrmann, H., Hertkorn,
531 N., Johnson, M., Liss, P. S., Quinn, P. K., Schartau, M., Soloviev, A., Stolle, C., Upstill-Goddard,
532 R. C., van Pinxteren, M., & Zäncker, B. (2017). The Ocean's Vital Skin: Toward an Integrated
533 Understanding of the Sea Surface Microlayer. *Front. Mar. Sci.*, 4(165).
534 <https://doi.org/10.3389/fmars.2017.00165>

- 535 Engel, A., Borchard, C., Piontek, J., Schulz, K. G., Riebesell, U., & Bellerby, R. (2013). CO₂ increases ¹⁴C
536 primary production in an Arctic plankton community. *Biogeosciences*, *10*(3), 1291-1308.
537 <https://doi.org/10.5194/bg-10-1291-2013>
- 538 Engel, A., & Galgani, L. (2016). The organic sea-surface microlayer in the upwelling region off the coast
539 of Peru and potential implications for air–sea exchange processes. *Biogeosciences*, *13*(4), 989-
540 1007. <https://doi.org/10.5194/bg-13-989-2016>
- 541 Engel, A., Sperling, M., Sun, C., Grosse, J., & Friedrichs, G. (2018). Organic Matter in the Surface
542 Microlayer: Insights From a Wind Wave Channel Experiment. *Front. Mar. Sci.*, *5*(182).
543 <https://doi.org/10.3389/fmars.2018.00182>
- 544 Engel, A., Thoms, S., Riebesell, U., Rochelle-Newall, E., & Zondervan, I. (2004). Polysaccharide
545 aggregation as a potential sink of marine dissolved organic carbon. *Nature*, *428*(6986), 929-932.
546 <https://doi.org/https://doi.org/10.1038/nature02453>
- 547 Frew, N. M. (1997). The role of organic films in air–sea gas exchange. In P. S. Liss & R. A. Duce (Eds.),
548 *The Sea Surface and Global Change* (pp. 121-172). Cambridge University Press.
549 <https://doi.org/DOI:10.1017/CBO9780511525025.006>
- 550 Galgani, L., & Engel, A. (2016). Changes in optical characteristics of surface microlayers hint to
551 photochemically and microbially mediated DOM turnover in the upwelling region off the coast of
552 Peru. *Biogeosciences*, *13*(8), 2453-2473. <https://doi.org/10.5194/bg-13-2453-2016>
- 553 Galgani, L., Engel, A., Rossi, C., Donati, A., & Loiselle, S. A. (2018). Polystyrene microplastics increase
554 microbial release of marine Chromophoric Dissolved Organic Matter in microcosm experiments.
555 *Sci. Rep.*, *8*(1), 14635. <https://doi.org/10.1038/s41598-018-32805-4>
- 556 Galgani, L., & Loiselle, S. A. (2019). Plastic Accumulation in the Sea Surface Microlayer: An Experiment-
557 Based Perspective for Future Studies. *Geosciences*, *9*.
558 <https://doi.org/https://doi.org/10.3390/geosciences9020066>
- 559 Galgani, L., & Loiselle, S. A. (2021). Plastic pollution impacts on marine carbon biogeochemistry. *Environ.*
560 *Pollut.*, *268*, 115598. <https://doi.org/https://doi.org/10.1016/j.envpol.2020.115598>
- 561 Galgani, L., Piontek, J., & Engel, A. (2016). Biopolymers form a gelatinous microlayer at the air-sea
562 interface when Arctic sea ice melts. *Sci. Rep.*, *6*, 29465.
- 563 Galgani, L., Stolle, C., Endres, S., Schulz, K. G., & Engel, A. (2014). Effects of ocean acidification on the
564 biogenic composition of the sea-surface microlayer: Results from a mesocosm study. *Journal of*
565 *Geophysical Research: Oceans*, *119*(11), 7911-7924. <https://doi.org/10.1002/2014jc010188>
- 566 Galgani, L., Tsapakis, M., Pitta, P., Tsiola, A., Tzempelikou, E., Kalantzi, I., Esposito, C., Loiselle, A.,
567 Tsotskou, A., Zivanovic, S., Dafnomili, E., Diliberto, S., Mylona, K., Magiopoulos, I., Zeri, C.,
568 Pitta, E., & Loiselle, S. A. (2019). Microplastics increase the marine production of particulate forms
569 of organic matter. *Environ. Res. Lett.* <https://doi.org/https://doi.org/10.1088/1748-9326/ab59ca>
- 570 Garrett, W. D. (1965). Collection of Slick-forming Materials from the Sea Surface. *Limnol. Oceanogr.*, *10*,
571 602-605. <https://doi.org/https://doi.org/10.4319/lo.1965.10.4.0602>
- 572 Geyer, R., Jambeck, J. R., & Law, K. L. (2017). Production, use, and fate of all plastics ever made. *Sci.*
573 *Adv.* *3*(7). <http://advances.sciencemag.org/content/3/7/e1700782.abstract>
- 574 Gupta, K. K., & Devi, D. (2020). Characteristics investigation on biofilm formation and biodegradation
575 activities of *Pseudomonas aeruginosa* strain ISJ14 colonizing low density polyethylene (LDPE)
576 surface. *Heliyon*, *6*(7), e04398. <https://doi.org/https://doi.org/10.1016/j.heliyon.2020.e04398>

- 577 Han, Y. N., Wei, M., Han, F., Fang, C., Wang, D., Zhong, Y. J., Guo, C. L., Shi, X. Y., Xie, Z. K., & Li,
578 F. M. (2020). Greater Biofilm Formation and Increased Biodegradation of Polyethylene Film by a
579 Microbial Consortium of *Arthrobacter* sp. and *Streptomyces* sp. *Microorganisms*, 8(12).
580 <https://doi.org/10.3390/microorganisms8121979>
- 581 Harlay, J., De Bodt, C., Engel, A., Jansen, S., d'ÄöHoop, Q., Piontek, J., Van Oostende, N., Groom, S.,
582 Sabbe, K., & Chou, L. (2009). Abundance and size distribution of transparent exopolymer particles
583 (TEP) in a coccolithophorid bloom in the northern Bay of Biscay. *Deep Sea Res. Part I Oceanogr.*
584 *Res. Pap.*, 56(8), 1251-1265. <https://doi.org/10.1016/j.dsr.2009.01.014>
- 585 Helms, J. R., Stubbins, A., Ritchie, J. D., Minor, E. C., Kieber, D. J., & Mopper, K. (2008). Absorption
586 spectral slopes and slope ratios as indicators of molecular weight, source, and photobleaching of
587 chromophoric dissolved organic matter. *Limnol. Oceanogr.*, 53(3), 955-969.
588 <https://doi.org/https://doi.org/10.4319/lo.2008.53.3.0955>
- 589 Hepach, H., Quack, B., Tegtmeier, S., Engel, A., Bracher, A., Fuhlbrügge, S., Galgani, L., Atlas, E. L.,
590 Lampel, J., Frieß, U., & Krüger, K. (2016). Biogenic halocarbons from the Peruvian upwelling
591 region as tropospheric halogen source. *Atmos. Chem. Phys.*, 16(18), 12219-12237.
592 <https://doi.org/10.5194/acp-16-12219-2016>
- 593 Hwang, J., Choi, D., Han, S., Jung, S. Y., Choi, J., & Hong, J. (2020). Potential toxicity of polystyrene
594 microplastic particles. *Sci. Rep.*, 10(1), 7391. <https://doi.org/10.1038/s41598-020-64464-9>
- 595 Jardillier, L., Zubkov, M. V., Pearman, J., & Scanlan, D. J. (2010). Significant CO₂ fixation by small
596 prymnesiophytes in the subtropical and tropical northeast Atlantic Ocean. *ISME J.*, 4(9), 1180-
597 1192. <https://doi.org/10.1038/ismej.2010.36>
- 598 Johnson, K. M., Dickson, A. G., Eiseid, G., Goyet, C., Guenther, P., Key, R. M., Millero, F. J., Purkerson,
599 D., Sabine, C. L., Schottle, R. G., Wallace, D. W. R., Wilke, R. J., & Winn, C. D. (1998).
600 Coulometric total carbon dioxide analysis for marine studies: assessment of the quality of total
601 inorganic carbon measurements made during the US Indian Ocean CO₂ Survey 1994–1996. *Mar.*
602 *Chem.*, 63(1), 21-37. [https://doi.org/https://doi.org/10.1016/S0304-4203\(98\)00048-6](https://doi.org/https://doi.org/10.1016/S0304-4203(98)00048-6)
- 603 Kvale, K., Prowe, A. E. F., Chien, C. T., Landolfi, A., & Oschlies, A. (2021). Zooplankton grazing of
604 microplastic can accelerate global loss of ocean oxygen. *Nat. Commun.* , 12(1), 2358.
605 <https://doi.org/10.1038/s41467-021-22554-w>
- 606 Lear, G., Kingsbury, J. M., Franchini, S., Gambarini, V., Maday, S. D. M., Wallbank, J. A., Weaver, L., &
607 Pantos, O. (2021). Plastics and the microbiome: impacts and solutions. *Environmental Microbiome*,
608 16(1), 2. <https://doi.org/10.1186/s40793-020-00371-w>
- 609 Lee, Y. K., Murphy, K. R., & Hur, J. (2020). Fluorescence Signatures of Dissolved Organic Matter Leached
610 from Microplastics: Polymers and Additives. *Environ. Sci. Technol.*, 54(19), 11905-11914.
611 <https://doi.org/10.1021/acs.est.0c00942>
- 612 Lindgren, J. F., Wilewska-Bien, M., Granhag, L., Andersson, K., & Eriksson, K. M. (2016). Discharges to
613 the Sea. In K. Andersson, S. Brynolf, J. F. Lindgren, & M. Wilewska-Bien (Eds.), *Shipping and*
614 *the Environment : Improving Environmental Performance in Marine Transportation* (pp. 125-168).
615 Springer Berlin Heidelberg. https://doi.org/10.1007/978-3-662-49045-7_4
- 616 Liss, P. S., & Duce, R. A. (2005). *The Sea Surface and Global Change*. Cambridge University Press.
617 <https://doi.org/doi:10.1017/CBO9780511525025>
- 618 Loiselle, S. A., Bracchini, L., Dattilo, A. M., Ricci, M., Tognazzi, A., Cózar, A., & Rossi, C. (2009). The
619 optical characterization of chromophoric dissolved organic matter using wavelength distribution of
620 absorption spectral slopes. *Limnol. Oceanogr.*, 54(2), 590-597.
621 <https://doi.org/10.4319/lo.2009.54.2.0590>

- 622 Lueker, T. J., Dickson, A. G., & Keeling, C. D. (2000). Ocean pCO₂ calculated from dissolved inorganic
623 carbon, alkalinity, and equations for K₁ and K₂: validation based on laboratory measurements of
624 CO₂ in gas and seawater at equilibrium. *Mar. Chem.*, 70(1), 105-119.
625 [https://doi.org/https://doi.org/10.1016/S0304-4203\(00\)00022-0](https://doi.org/https://doi.org/10.1016/S0304-4203(00)00022-0)
- 626 Mari, X. (1999). Carbon content and C:N ratio of transparent exopolymeric particles (TEP) produced by
627 bubbling exudates of diatoms. *Mar. Ecol. Prog. Ser.*, 183, 59-71.
628 <https://doi.org/10.3354/meps183059>
- 629 Mari, X., & Burd, A. (1998). Seasonal size spectra of transparent exopolymeric particles (TEP) in a coastal
630 sea and comparison with those predicted using coagulation theory. *Mar. Ecol. Prog. Ser.*, 163, 63-
631 76. <https://doi.org/10.3354/meps163063>
- 632 Mari, X., & Kiørboe, T. (1996). Abundance, size distribution and bacterial colonization of transparent
633 exopolymeric particles (TEP) during spring in the Kattegat. *J. Plankton Res.*, 18(6), 969-986.
634 <https://doi.org/10.1093/plankt/18.6.969>
- 635 Marie, D., Partensky, F., Jacquet, S., & Vaulot, D. (1997). Enumeration and Cell Cycle Analysis of Natural
636 Populations of Marine Picoplankton by Flow Cytometry Using the Nucleic Acid Stain SYBR Green
637 I. *Appl. Environ. Microbiol.*, 63(1), 186-193. [https://doi.org/https://doi.org/10.1128/aem.63.1.186-
638 193.1997](https://doi.org/https://doi.org/10.1128/aem.63.1.186-193.1997)
- 639 Michels, J., Stippkugel, A., Lenz, M., Wirtz, K., & Engel, A. (2018). Rapid aggregation of biofilm-covered
640 microplastics with marine biogenic particles. *Proc. Royal Soc. B.*
641 <https://doi.org/http://dx.doi.org/10.1098/rspb.2018.1203>
- 642 Mintrop, L., Pérez, F. F., González-Dávila, M., Santana-Casiano, J. M., & Körtzinger, A. (2000). Alkalinity
643 determination by potentiometry: intercalibration using three different methods. *Cienc. Mar.*, 26(1),
644 23-27. <https://doi.org/http://dx.doi.org/10.7773/cm.v26i1.573>
- 645 Mustaffa, N. I. H., Ribas-Ribas, M., Banko-Kubis, H. M., & Wurl, O. (2020). Global reduction of in situ
646 CO₂ transfer velocity by natural surfactants in the sea-surface microlayer. *Proc. Math. Phys. Eng.*
647 *Sci.*, 476(2234), 20190763. <https://doi.org/10.1098/rspa.2019.0763>
- 648 Orr, J. C., Epitalon, J. M., & Gattuso, J. P. (2015). Comparison of ten packages that compute ocean
649 carbonate chemistry. *Biogeosciences*, 12(5), 1483-1510. <https://doi.org/10.5194/bg-12-1483-2015>
- 650 Ortega-Retuerta, E., Mazuecos, I. P., Reche, I., Gasol, J. M., Álvarez-Salgado, X. A., Álvarez, M., Montero,
651 M. F., & Arístegui, J. (2019). Transparent exopolymer particle (TEP) distribution and in situ
652 prokaryotic generation across the deep Mediterranean Sea and nearby North East Atlantic Ocean.
653 *Prog. Oceanogr.*, 173, 180-191. <https://doi.org/https://doi.org/10.1016/j.pocean.2019.03.002>
- 654 Pabortsava, K., & Lampitt, R. S. (2020). High concentrations of plastic hidden beneath the surface of the
655 Atlantic Ocean. *Nat. Comm.*, 11(1), 4073. <https://doi.org/10.1038/s41467-020-17932-9>
- 656 Paerl, H. W. (1975). Microbial attachment to particles in marine and freshwater ecosystems. *Microb. Ecol.*,
657 2(1), 73-83. <https://doi.org/10.1007/bf02010382>
- 658 Pierrot, D., Lewis, E., & Wallace, D. W. R. (2006). MS Excel Program Developed for CO₂ System
659 Calculations. ORNL/CDIAC-105a. Carbon Dioxide Information Analysis Center, Oak Ridge
660 National Laboratory, U.S. Department of Energy, Oak Ridge, Tennessee. doi:
661 10.3334/CDIAC/otg.CO2SYS_XLS_CDIAC105a
- 662 Pitta, P., Nejstgaard, J. C., Tsagaraki, T. M., Zervoudaki, S., Egge, J. K., Frangoulis, C., Lagaria, A.,
663 Magiopoulos, I., Psarra, S., Sandaa, R.-A., Skjoldal, E. F., Tanaka, T., Thyrraug, R., & Thingstad,
664 T. F. (2016). Confirming the “Rapid phosphorus transfer from microorganisms to mesozooplankton

- 665 in the Eastern Mediterranean Sea” scenario through a mesocosm experiment. *J. Plankton Res.*,
666 38(3), 502-521. <https://doi.org/10.1093/plankt/fbw010>
- 667 Rahav, E., Shun-Yan, C., Cui, G., Liu, H., Tsagaraki, T. M., Giannakourou, A., Tsiola, A., Psarra, S.,
668 Lagaria, A., Mulholland, M. R., Stathopoulou, E., Paraskevi, P., Herut, B., & Berman-Frank, I.
669 (2016). Evaluating the Impact of Atmospheric Depositions on Springtime Dinitrogen Fixation in
670 the Cretan Sea (Eastern Mediterranean)—A Mesocosm Approach. *Front. Mar. Sci.*, 3.
671 <https://doi.org/10.3389/fmars.2016.00180>
- 672 Rahlff, J., Stolle, C., Giebel, H.-A., Brinkhoff, T., Ribas-Ribas, M., Hodapp, D., & Wurl, O. (2017). High
673 wind speeds prevent formation of a distinct bacterioneuston community in the sea-surface
674 microlayer. *FEMS Microbiol. Ecol.*, 93(5), fix041-fix041. <https://doi.org/10.1093/femsec/fix041>
- 675 Rahlff, J., Stolle, C., Giebel, H.-A., Ribas-Ribas, M., Damgaard, L. R., & Wurl, O. (2019). Oxygen Profiles
676 Across the Sea-Surface Microlayer—Effects of Diffusion and Biological Activity [Original
677 Research]. *Front. Mar. Sci.*, 6(11). <https://doi.org/10.3389/fmars.2019.00011>
- 678 Royer, S.-J., Ferrón, S., Wilson, S. T., & Karl, D. M. (2018). Production of methane and ethylene from
679 plastic in the environment. *PLOS ONE*, 13(8), e0200574.
680 <https://doi.org/10.1371/journal.pone.0200574>
- 681 Santschi, P. H., Chin, W.-C., Quigg, A., Xu, C., Kamalanathan, M., Lin, P., & Shiu, R.-F. (2021). Marine
682 Gel Interactions with Hydrophilic and Hydrophobic Pollutants. *Gels*, 7(3), 83.
683 <https://www.mdpi.com/2310-2861/7/3/83>
- 684 Savoca, M. S., Wohlfeil, M. E., Ebeler, S. E., & Nevitt, G. A. (2016). Marine plastic debris emits a keystone
685 infochemical for olfactory foraging seabirds. *Science Advances*, 2(11), e1600395.
686 <https://doi.org/doi:10.1126/sciadv.1600395>
- 687 Shen, M., Ye, S., Zeng, G., Zhang, Y., Xing, L., Tang, W., Wen, X., & Liu, S. (2020). Can microplastics
688 pose a threat to ocean carbon sequestration? *Mar. Pollut. Bull.*, 150, 110712.
689 <https://doi.org/https://doi.org/10.1016/j.marpolbul.2019.110712>
- 690 Shiu, R.-F., Vazquez, C. I., Chiang, C.-Y., Chiu, M.-H., Chen, C.-S., Ni, C.-W., Gong, G.-C., Quigg, A.,
691 Santschi, P. H., & Chin, W.-C. (2020). Nano- and microplastics trigger secretion of protein-rich
692 extracellular polymeric substances from phytoplankton. *Sci.Total. Environ.*, 748, 141469.
693 <https://doi.org/https://doi.org/10.1016/j.scitotenv.2020.141469>
- 694 Su, X., Yang, L., Yang, K., Tang, Y., Wen, T., Wang, Y., Rillig, M. C., Rohe, L., Pan, J., Li, H., & Zhu,
695 Y.-g. (2022). Estuarine plastisphere as an overlooked source of N₂O production. *Nature*
696 *Communications*, 13(1), 3884. <https://doi.org/10.1038/s41467-022-31584-x>
- 697 Thompson, J. B. (2000). Microbial Whittings. In R. E. Riding & S. M. Awramik (Eds.), *Microbial Sediments*
698 (pp. 250-260). Springer Berlin Heidelberg. https://doi.org/10.1007/978-3-662-04036-2_27
- 699 Thompson, J. B., Schultze-Lam, S., Beveridge, T. J., & Des Marais, D. J. (1997). Whiting events: biogenic
700 origin due to the photosynthetic activity of cyanobacterial picoplankton. *Limnol. Oceanogr.*, 42(1),
701 133-141. <https://doi.org/10.4319/lo.1997.42.1.0133>
- 702 Thornton, D. C. O. (2018). Coomassie Stainable Particles (CSP): Protein Containing Exopolymer Particles
703 in the Ocean. *Front. Mar. Sci.*, 5(206). <https://doi.org/10.3389/fmars.2018.00206>
- 704 Thornton, D. C. O., Brooks, S. D., & Chen, J. (2016). Protein and Carbohydrate Exopolymer Particles in
705 the Sea Surface Microlayer (SML). *Front. Mar. Sci.*, 3(135).
706 <https://doi.org/10.3389/fmars.2016.00135>
- 707 Tsiola, A., Pitta, P., Callol, A. J., Kagiorgi, M., Kalantzi, I., Mylona, K., Santi, I., Toncelli, C., Pergantis,
708 S., & Tsapakis, M. (2017). The impact of silver nanoparticles on marine plankton dynamics:

- 709 Dependence on coating, size and concentration. *Sci.Total. Environ.*, 601-602, 1838-1848.
710 <https://doi.org/https://doi.org/10.1016/j.scitotenv.2017.06.042>
- 711 Tsiola, A., Tsagaraki, T. M., Giannakourou, A., Nikolioudakis, N., Yücel, N., Herut, B., & Pitta, P. (2017).
712 Bacterial Growth and Mortality after Deposition of Saharan Dust and Mixed Aerosols in the
713 Eastern Mediterranean Sea: A Mesocosm Experiment. *Frontiers in Marine Science*, 3.
714 <https://doi.org/10.3389/fmars.2016.00281>
- 715 Upstill-Goddard, R. C. (2006). Air–sea gas exchange in the coastal zone. *Estuarine, Coastal and Shelf*
716 *Science*, 70(3), 388-404. <https://doi.org/https://doi.org/10.1016/j.ecss.2006.05.043>
- 717 Verdugo, P. (2012). Marine Microgels. *Annual Review of Marine Science*, 4(1), 375-400.
718 <https://doi.org/doi:10.1146/annurev-marine-120709-142759>
- 719 Wurl, O., Ekau, W., Landing, W. M., & Zappa, C. J. (2017). Sea surface microlayer in a changing ocean –
720 A perspective. *Elementa-Sci. Anthropol.*, 5(31). <https://doi.org/http://doi.org/10.1525/elementa.228>
- 721 Wurl, O., & Holmes, M. (2008). The gelatinous nature of the sea-surface microlayer. *Mar. Chem.*, 110(1-
722 2), 89-97. <https://doi.org/10.1016/j.marchem.2008.02.009>
- 723 Wurl, O., Stolle, C., Van Thuoc, C., The Thu, P., & Mari, X. (2016). Biofilm-like properties of the sea
724 surface and predicted effects on air–sea CO₂ exchange. *Prog. Oceanogr.*, 144, 15-24.
725 <https://doi.org/https://doi.org/10.1016/j.pocean.2016.03.002>
- 726 Yue, W.-z., Sun, C.-c., Shi, P., Engel, A., Wang, Y.-s., & He, W.-H. (2018). Effect of temperature on the
727 accumulation of marine biogenic gels in the surface microlayer near the outlet of nuclear power
728 plants and adjacent areas in the Daya Bay, China. *PLOS ONE*, 13(6), e0198735.
729 <https://doi.org/10.1371/journal.pone.0198735>
- 730 Zäncker, B., Bracher, A., Röttgers, R., & Engel, A. (2017). Variations of the Organic Matter Composition
731 in the Sea Surface Microlayer: A Comparison between Open Ocean, Coastal, and Upwelling Sites
732 Off the Peruvian Coast. *Front. Microbiol.*, 8(2369). <https://doi.org/10.3389/fmicb.2017.02369>
- 733 Zäncker, B., Cunliffe, M., & Engel, A. (2018). Bacterial Community Composition in the Sea Surface
734 Microlayer Off the Peruvian Coast. *Front. Microbiol.*, 9(2699).
735 <https://doi.org/10.3389/fmicb.2018.02699>
- 736 Zettler, E. R., Mincer, T. J., & Amaral-Zettler, L. A. (2013). Life in the “Plastisphere”: Microbial
737 Communities on Plastic Marine Debris. *Environ. Sci. Technol.*, 47(13), 7137-7146.
738 <https://doi.org/10.1021/es401288x>
- 739 Zhao, S., Zettler, E. R., Amaral-Zettler, L. A., & Mincer, T. J. (2021). Microbial carrying capacity and
740 carbon biomass of plastic marine debris. *ISME J.*, 15(1), 67-77. <https://doi.org/10.1038/s41396-020-00756-2>

742

743

744

Parameter / unit	ANOVA F	P value	Test	Mean Control	Mean MP	SE of difference	% of difference MP-C
TA (ULW) [$\mu\text{mol Kg}^{-1}$]	F (11, 43) = 3.138	0.0035	REML	2625.0	2631.0	1.77	0.2
DIC (ULW)* [$\mu\text{mol Kg}^{-1}$]	F (11, 43) = 9.482	<0.0001	REML (time)	2283.0	2286.0	0.94	0.1
pCO ₂ (ULW) [μatm]	F (11, 43) = 4.279	0.0003	REML	420.20	409.10	3.78	-2.6
pH (ULW)	F (11, 43) = 4.552	0.0001	REML	8.064	8.075	0.00	0.14
CO ₃ ²⁻ [$\mu\text{mol Kg}^{-1}$]	F (11, 43) = 2.814	0.0075	REML	243.10	246.20	1.09	1.3
CO ₂ [$\mu\text{mol Kg}^{-1}$]	F (11, 43) = 3.845	0.0007	REML	13.09	12.78	0.11	-2.4
HCO ₃ ⁻ * [$\mu\text{mol Kg}^{-1}$]	F (11, 43) = 2.991	0.0049	REML (time)	2027.0	2027.0	1.27	0.0
TEP (SML)* [10 ⁶ mm ² L ⁻¹]	F (6, 24) = 3.573	0.0113	RM ANOVA (time)	493.50	641.70	60.30	30.0
TEP (SML) [10 ⁶ particles L ⁻¹]	F (6, 24) = 2.836	0.0312	RM ANOVA	61.06	77.44	5.53	26.8
TEP-C (SML)* [$\mu\text{g Carbon L}^{-1}$]	F (6, 24) = 3.868	0.0077	RM ANOVA (time)	324.30	424.80	49.43	31.0
CSP (SML) [10 ⁶ mm ² L ⁻¹]	F (6, 24) = 4.932	0.002	RM ANOVA	682.30	854.50	55.72	25.2
CSP (SML) [10 ⁶ particles L ⁻¹]	F (6, 24) = 4.870	0.0022	RM ANOVA	75.93	113.57	8.24	49.6
Syn (SML) [10 ⁶ cells L ⁻¹]	F (5, 20) = 4.417	0.0071	RM ANOVA	64.28	65.15	3.01	1.4
HDNA-Syn (SML)* [10 ⁶ cells L ⁻¹]	F (5, 20) = 21.54	<0.0001	RM ANOVA (time)	1.46	1.81	0.22	23.5
LDNA-Syn (SML) [10 ⁶ cells L ⁻¹]	F (5, 20) = 4.947	0.0041	RM ANOVA	62.85	63.35	2.84	0.8
S ₂₇₅₋₂₉₅ (SML) [nm ⁻¹]	F (6, 24) = 3.146	0.0203	RM ANOVA	0.01	0.01	0.00	4.1
a(355) nm (SML) [m ⁻¹]	F (6, 24) = 1.090	0.3963	RM ANOVA	3.89	3.90	0.05	0.1
DOC (SML) [mg L ⁻¹]	F (1, 4) = 0.02581	0.6927	REML	1.23	1.21	0.11	-1.5
H. Bacteria (SML)* [10 ⁸ cells L ⁻¹]	F (6, 21) = 9.815	<0.0001	REML (time)	4.46	4.31	0.42	-3.4
HDNA H. Bacteria (SML)* [10 ⁸ cells L ⁻¹]	F (6, 21) = 10.40	<0.0001	REML (time)	3.13	2.98	0.39	-4.8
LDNA H. Bacteria (SML)* [10 ⁸ cells L ⁻¹]	F (6, 21) = 13.07	<0.0001	REML (time)	1.27	1.34	0.05	5.3

746 Table 1 Repeated Measures Two-Way ANOVA or Mixed effects Model (REML) Analysis Table where the
747 fixed factor is the "treatment" (MP/ no MP) and the random effect is the time. Results are shown as the
748 interaction of "treatment x time" unless otherwise noted. Replicates within the treatments have been

749 assumed having equal variability of differences. (*) indicates differences that have been observed between
 750 MP and Control in time. Significant differences are accepted for $p < 0.05$.

751

$S_{275-295}$ [nm ⁻¹], SML	pCO_2 [μatm], ULW	DOC [mg L ⁻¹], SML	Heterotrophic bacteria [cells L ⁻¹], SML	CSP [mm ² L ⁻¹], SML	<i>Synechococcus</i> [cells L ⁻¹], SML
Spearman r	0.51	0.45	0.34	0.35	-0.61
p	< 0.001	0.004	0.032	0.024	< 0.0001
n	42	39	39	42	39

752 Table 2 Spearman r correlation table reporting significant correlations of Spectral Slope in the SML
 753 between 275 and 295 nm ($S_{275-295}$) to pCO_2 (ulw), and DOC, Heterotrophic Bacteria, CSP and *Synechococcus*
 754 cells in the SML.

755

756

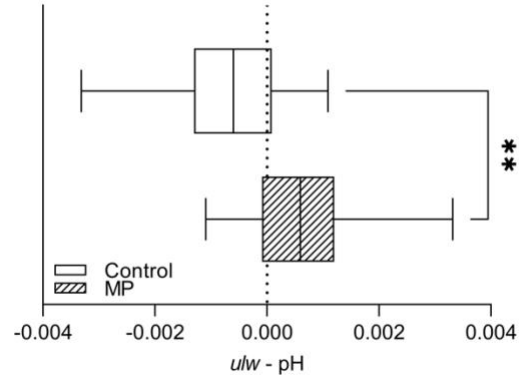
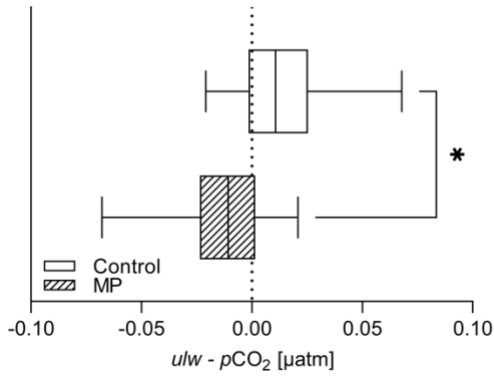
757

<i>Synechococcus</i> [cells L ⁻¹]	TA [$\mu mol Kg^{-1}$]	DIC [$\mu mol Kg^{-1}$]	pH in	pCO_2 [μatm]	CO_3^{2-} [$\mu mol Kg^{-1}$]	CO_2 [$\mu mol Kg^{-1}$]
<i>Spearman r</i>	0.531	0.429	0.344	-0.33	0.336	-0.321
<i>p</i>	0.000557	0.00664	0.0323	0.0406	0.0368	0.0465
<i>n</i>	39	39	39	39	39	39
TEP-C [$\mu g C L^{-1}$]	TA [$\mu mol Kg^{-1}$]	DIC [$\mu mol Kg^{-1}$]	pH in	pCO_2 [μatm]	CO_3^{2-} [$\mu mol Kg^{-1}$]	CO_2 [$\mu mol Kg^{-1}$]
<i>Spearman r</i>	0.511	0.273	0.434	-0.433	0.481	-0.437
<i>p</i>	0.000592	0.0798	0.00428	0.00438	0.00137	0.00395
<i>n</i>	42	42	42	42	42	42

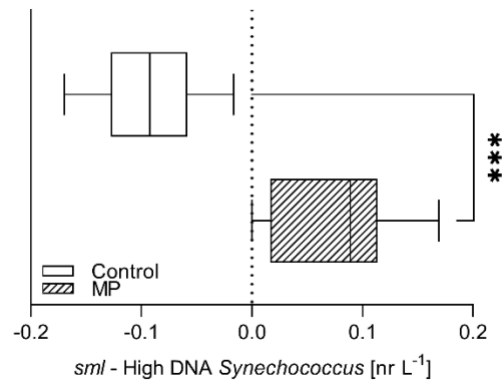
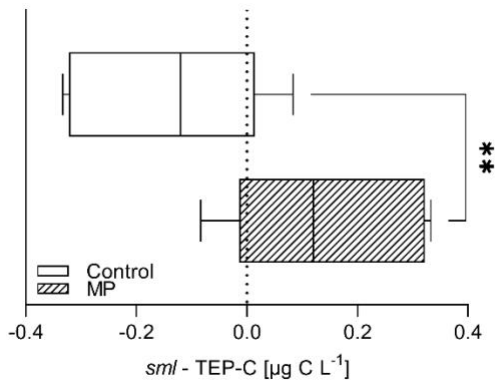
758 Table 3 Spearman r correlation table between *Synechococcus* and TEP-Carbon (TEP-C) in the SML to
 759 Total Alkalinity (TA), DIC, pH, pCO_2 , CO_3^{2-} and CO_2 in the bulk water of the mesocosms.

760

761

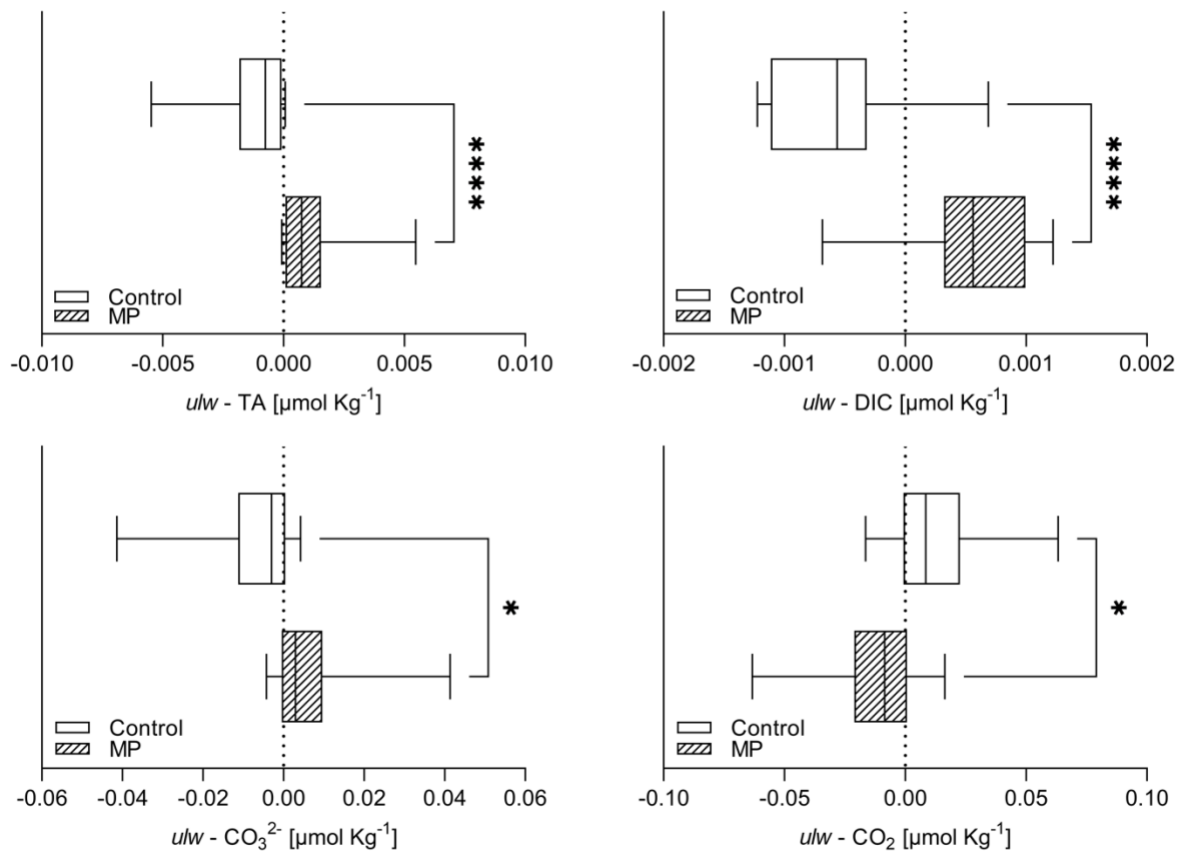


762



763

764 *Figure 1* Boxplots represent normalized anomalies and 5-95 percentiles for underlying water (ulw) pCO₂,
 765 pH, and for sea-surface microlayer (SML) TEP-carbon (TEP-C) and High DNA containing *Synechococcus*
 766 cells. Stars indicate the level of significance in the differences between control and MP mesocosms based
 767 on Mann-Whitney tests on normalized anomalies (pCO₂, pH, and unpaired t-tests (TEP-C and High DNA
 768 *Synechococcus* cells).



769

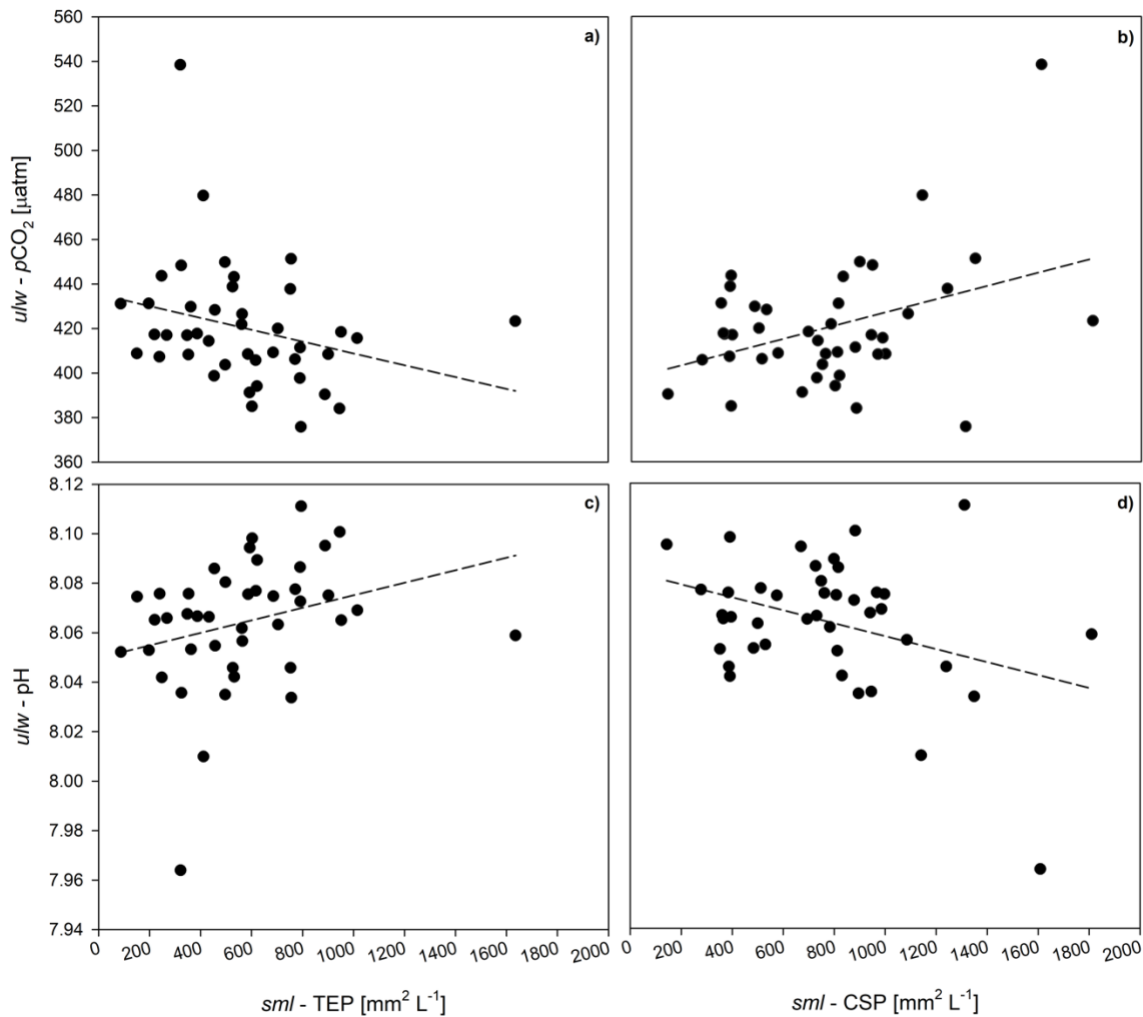
770 *Figure 2 Boxplots that represent normalized anomalies and 5-95 percentiles for underlying water (ulw)*

771 *Total Alkalinity (TA), Dissolved Inorganic Carbon (DIC), CO_3^{2-} and CO_2 . Stars indicate the level of*

772 *significance in the differences between control and MP mesocosms based on Mann-Whitney test (TA,*

773 *CO_3^{2-} and CO_2) and unpaired t-test (DIC).*

774



775

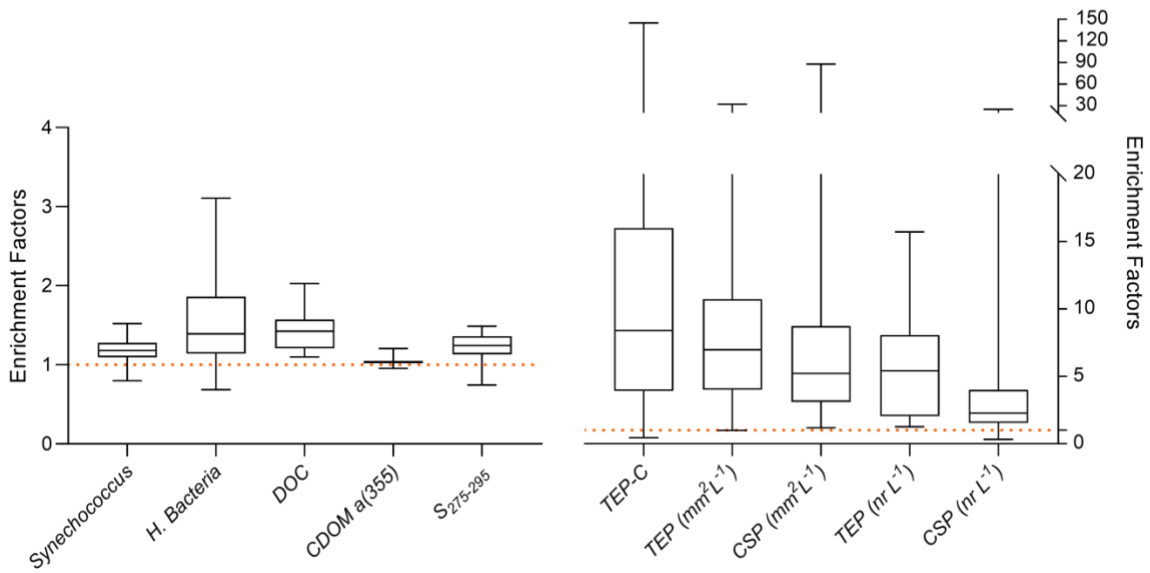
776 *Figure 3 Panels a, b: Multiple linear regression between pCO₂ in the underlying water (ulw) and the*
 777 *presence of marine gels (TEP, a; CSP, b) in the sea-surface microlayer (SML).*

778 *Panels c, d: Multiple linear regression between pH in the underlying water (ulw) on marine gels (TEP, c;*
 779 *CSP, d) in the sea-surface microlayer (SML) according to the equation:*

780 $pH = 8.070 + 0.0000471 \text{ TEP [mm}^2 \text{ L}^{-1}] - 0.0000421 \text{ CSP [mm}^2 \text{ L}^{-1}].$

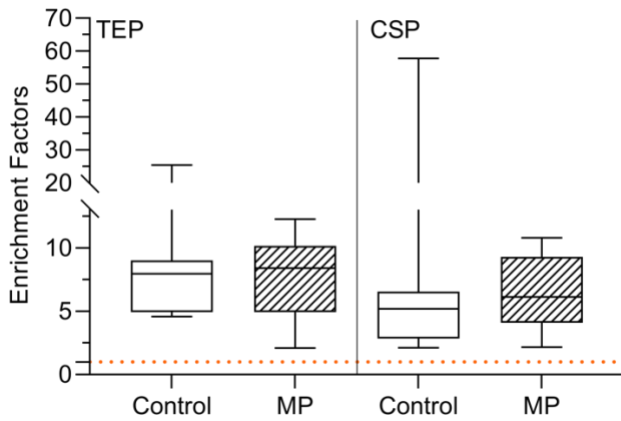
781 *For the regression, $R^2 = 0.365$, $p < 0.001$, $F = 11.22$, $DF = 2$. All panels have been displayed separately to*
 782 *better visualize the trends.*

783



784

785 *Figure 4 Enrichment factors for SML parameters compared to underlying water considering both*
 786 *treatments. The dashed orange line is set on both graphs to EF = 1 which means no real differences*
 787 *between SML concentration and underlying water concentration.*



788

789 *Figure 5 Enrichment Factors for the concentration of marine gels TEP and CSP, expressed as area mm²*
 790 *L⁻¹, between Control and MP mesocosms. The orange dashed line indicates EF = 1. Repeated Measures*
 791 *two-way ANOVA tests (Table S4) have evidenced significant differences between EFs.*

792

793